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2 purpose

3 **Running title:** Tools for understanding fish populations

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36

37 **Abstract**

38 Wild fish populations are currently experiencing unprecedented pressures, which are  
39 projected to intensify in the coming decades. Developing a thorough understanding  
40 of the influences of both biotic and abiotic factors on fish populations is a salient  
41 issue in contemporary fish conservation and management. During the 50<sup>th</sup>  
42 Anniversary of the Fisheries Society of the British Isles, University of Exeter, 2017,  
43 scientists from diverse research backgrounds gathered to discuss key topics under  
44 the broad umbrella of 'Understanding Fish Populations'. Below, the output of one  
45 such discussion group is detailed, focusing on tools used to investigate natural fish  
46 populations. Five main groups of approaches were identified: (i) Tagging and  
47 telemetry; (ii) Molecular tools; (iii) Survey tools; (iv) Statistical and modelling tools;  
48 and (v) Tissue analyses. The appraisal covered current challenges and potential  
49 solutions for each of these topics. In addition, three key themes were identified as  
50 applicable across all tool-based applications. These included data management,  
51 public engagement, and fisheries policy and governance. The continued innovation  
52 of tools and capacity to integrate interdisciplinary approaches into the future  
53 assessment and management of fish populations is highlighted as an important focus  
54 for the next 50 years of fisheries research.

55 Key words: archaeology, genetics, modelling, surveys, stable isotopes, telemetry

56

## 57 **Introduction**

58           Approximately 30% of fish species have been overexploited (FAO, 2014),  
59 representing significant losses to biodiversity, ecosystem services and  
60 socioeconomic contributions (Worm et al., 2006). In light of the increasing challenges  
61 presented by climate change and other natural and anthropogenic stressors (Gordon  
62 et al., 2017), an improved understanding of fish populations is critical to facilitate  
63 effective management and conservation initiatives. During the summer of 2017, the  
64 Fisheries Society of the British Isles (FSBI) held its 50<sup>th</sup> Anniversary Symposium  
65 under the broad umbrella of ‘Understanding Fish Populations’. To highlight key  
66 knowledge gaps and opportunities, we detail the outcome of a working group  
67 convened at the symposium, which was tasked with considering the theme of ‘Tools  
68 for understanding fish populations’. The scope of the discussion spanned diverse  
69 areas including spatial ecology and migration patterns, genetics and evolutionary  
70 biology, physiology, trophic ecology, and developmental and population biology. In  
71 this article, we consider major advancements in the use of tools across broad areas  
72 of fish biology, and identify knowledge gaps and potential solutions in each area in  
73 order to guide and inform future research, to better understand and protect wild fish  
74 populations.

## 75 **Tagging and telemetry**

76 A significant problem hampering the study of marine fish is that of determining their  
77 geographical locations at fine scales, over long durations and, in particular, for  
78 benthic species. Tagging and telemetry involves the application of external and or  
79 internal tags or devices to manually or passively track fish movement (Cooke et al.,  
80 2013). Both forms can be particularly challenging in the marine environment, though

81 manual tracking can work at feeding grounds and at spawning aggregations (e.g.  
82 Murchie et al., 2015), while passive tracking has valuable applications along known  
83 migration routes (Dahlgren et al., 2016), for example, as anadromous/catadromous  
84 species migrate in and out of river estuaries (e.g. Lauridsen et al., 2017). Suites of  
85 tools exist for such tasks (e.g. acoustic transmitters, PIT and Floy™ tags, radio,  
86 archival, etc.) and have been routinely used to understand the spatial ecology of a  
87 range of fish taxa (Bograd et al., 2010). With technological improvements in tags and  
88 tracking equipment, the field has grown vastly in recent decades (see reviews by  
89 Pine et al., 2003; Bograd et al., 2010; Jepsen et al., 2015). We briefly highlight some  
90 of the tags and telemetry options commonly used by researchers along with a  
91 discussion of some of the limitations and challenges associated with these tools.

92 Data storage tags (DSTs), which collect data on both the internal and/or external  
93 environments of fish are the only method available to assess internal states (e.g.  
94 bioenergetics, Cooke et al., 2016). However, DSTs currently only provide information  
95 on the environment experienced by the tagged fish if the tag is recovered, meaning  
96 these data are lost if fish recapture rates are low, often the case in fish tagging  
97 surveys. Communication History Acoustic Tags (so called 'CHATs'), which transmit  
98 data to nearby transponder receivers are a promising alternative. Since there have  
99 been relatively few uses of this tag type (Voegeli et al., 2001; Hight & Lowe, 2007),  
100 there is potential for development in this area.

101 Pop-up satellite archival tags (PSATs), which detach from the tagged fish after some  
102 time at sea and transmit telemetry data to overpassing satellites, are currently limited  
103 in terms of hardware, software and satellite reception. PSATs are large, so are  
104 limited in use for larger, often highly migratory individuals, and may also affect fish

105 behaviour (Methling et al., 2011). Additionally, battery failure, antenna damage, or  
106 mechanical failure may limit registration or transmission of data (Musyl et al., 2011).  
107 PSAT technology is relatively new, so future reductions in size and weight and also  
108 improvement in reliability can be expected. In terms of software, PSATs currently  
109 only transmit limited amounts of data due to transmission costs and the short time  
110 that the receiving satellite is above the horizon. Future software development is  
111 required to reduce transmission costs, optimise data transmission and provide more  
112 flexibility for users to tailor controls to provide higher resolution data at the desired  
113 temporal scale. An increase in the number of satellite platforms that can receive  
114 PSAT data would help to improve reception issues. Interference on frequencies  
115 selected for tags at certain geographical locations (see Musyl et al., 2011) also  
116 requires consideration.

117 Acoustic telemetry offers autonomous, continuous monitoring (Heupel et al., 2006)  
118 and has the potential to significantly enhance our understanding of marine predator  
119 habitat use, activity patterns and resource partitioning (Hussey et al., 2015). Acoustic  
120 arrays have been used in many studies in elucidating fish movements (e.g.  
121 Papastamatiou et al., 2013; Lea et al., 2016), and transmitters have been used more  
122 innovatively to measure trophic interactions (Halfyard et al., 2017). Issues remain  
123 however, for example, in the significant cost and effort involved in deploying and  
124 maintaining acoustic arrays.

125 Organisations such as the Ocean Tracking Network (Whoriskey et al., 2015), (OTN;  
126 <http://oceantrackingnetwork.org>) and the Australian Animal Tracking Network both  
127 maintain acoustic infra-structure in the form of deployed receivers (arrays or curtains)  
128 in key ecological areas into which researchers are free to release tagged animals.

129 These initiatives substantially reduce the cost and risk associated with acoustic  
130 tracking projects and similar approaches can be applied globally. Furthermore,  
131 integration of data standardised repositories along with a comprehensive set of  
132 analytical tools to ensure rapid and sophisticated analysis of acoustic array data  
133 would lead to new insights into the spatial ecology of fish. Further technological  
134 developments such as the use of AUVs to perform routine data download operations,  
135 or even complement fixed acoustic receivers (Davis et al., 2016), will make acoustic  
136 telemetry increasingly affordable and accessible to more researchers. Continued  
137 collaborations with established regional and international tracking networks, together  
138 with the ever-increasing sophistication, miniaturisation, durability and cost reduction  
139 of tags promises an increasingly important role for acoustic telemetry in our  
140 understanding of fish ecology.

## 141 **Molecular tools**

### 142 *Population genetics and genomics*

143 Using genetic tools to understand fish genetic diversity and population structure has  
144 wide-ranging applications for evolutionary biology, and the conservation and  
145 management of fish stocks. Until recently, molecular techniques such as  
146 mitochondrial sequencing and the analysis of microsatellite loci have been used most  
147 commonly to explore intra-specific variation in fish and many other organisms (e.g.  
148 Ferguson & Danzmann 1998; Chistiakov et al., 2006; Abdul-Muneer, 2014). More  
149 recently, however, the increased availability and cost efficiency of high-throughput  
150 sequencing, which is capable of producing millions of sequencing reads (e.g.  
151 RADseq, RNAseq), has revolutionised the fields of population and conservation

152 genetics (Allendorf et al., 2010). It is however important to understand what extra  
153 information high-throughput sequencing data can provide, the biases involved in  
154 study design and data generation, and also how its usage might be optimised. Here,  
155 we seek to identify knowledge gaps in the field of fish population genetics, and  
156 contemplate how this area of research may evolve in the future.

157 Attaining high quality, clean DNA for large numbers of individuals is paramount for  
158 downstream sequencing processes, but in some cases can be challenging. Biological  
159 samples can often be compromised during sampling or transport, potentially  
160 rendering field efforts futile. Population genetic studies on fish frequently require  
161 sampling from river transects or remote locations at sea, and so portable laboratories  
162 for sampling, storing and extracting DNA would be welcomed. At the same time,  
163 emerging technologies, e.g. the MinION USB sequencer  
164 (<https://nanoporetech.com/products/minion>), have the potential to revolutionise when  
165 and where genetic data can be generated. Most new technologies are currently  
166 restricted to sequencing small genomes, such as those of bacteria, but with on-going  
167 improvements, these technologies open up the possibility of being able to sequence  
168 DNA in real-time in the field (Hayden, 2015). Recently, the MinION technology has  
169 started to be used in hybrid assemblies with Illumina short reads (Austin et al., 2017)  
170 and *de novo* eukaryotic genomes (including fish) are in progress (Jansen et al.,  
171 2017).

172 Alongside population genetic studies, research based on whole genome data is  
173 emerging, and the genomes of several commercially important species have now  
174 been published (e.g. Atlantic cod (*Gadus morhua*), Star et al., 2011; Atlantic salmon  
175 (*Salmo salar*), Lien et al., 2016). However, while the ever-reducing cost of whole

176 genome sequencing provides opportunities to sequence and publish more fish  
177 genomes, in our view, the key priority is not simply publishing genomes, but also  
178 high-quality genome annotation. Gene annotation and accurate knowledge of the  
179 function of different identified regions is of extreme importance if genomic tools are to  
180 be used reliably in conservation and management (Ekblom & Wolf, 2014). Therefore,  
181 projects such as the 'Functional Annotation of All Salmonid Genomes' (Macqueen et  
182 al., 2017) should be encouraged and developed. It is also important not to  
183 underestimate or neglect the computing power and bioinformatics expertise required  
184 to produce high quality genome scaffolds and annotations, and also to recognise and  
185 account for biases in next generation sequencing data (see Benestan et al., 2016).

186 Furthermore, population genetic approaches are usually focused on a single species.  
187 Consequently, there is a mismatch between studies of a single species genotyped at  
188 high resolution, but generally at small spatial scales (e.g. population genetics, often  
189 using hundreds to thousands of markers through GBS or GWAS) and studies of  
190 multiple species at larger spatial scales but using lower resolution markers (e.g.  
191 phylogeography or biodiversity assessments using metabarcoding or mtDNA  
192 sequencing). Nonetheless, the widespread application of molecular resources has  
193 led to the accumulation of rich datasets across a broad range of species,  
194 geographical regions and time periods (Blanchet et al., 2017). Accordingly, we  
195 anticipate that this aggregation of data may allow the underlying processes that drive  
196 genetic variability across these regions and times to be revealed, enabling a broader  
197 testing of theories in population genetics and evolution (Ellegren & Galtier, 2016;  
198 Pauls et al., 2014).

199 Such studies will require combining high genetic resolution markers across large  
200 spatial scales, which is a non-trivial task, especially when dealing with non-model  
201 species. Three challenges arise in such cases: firstly, the financial investment  
202 required to obtain reliable datasets for several species remains significant. Despite  
203 reductions in sequencing costs, it may be financially sensible to rely on more  
204 classical markers such as microsatellites or small subsets of single nucleotide  
205 polymorphisms (SNPs). Secondly, there is a need for a standardised framework in  
206 order to make datasets comparable across different species and regions. This  
207 standardisation must occur when collecting samples, characterising markers (e.g.  
208 Ellis et al., 2011; Helyar et al., 2011) and during the subsequent data analysis to  
209 streamline user choices (Paris et al., 2017), which may bias the biological  
210 interpretation of data, see Rodríguez-Ezpeleta et al. (2016). It is therefore important  
211 that researchers use common methods to isolate and characterise markers for entire  
212 sets of focal species, and/or provide full access to detailed analyses when datasets  
213 are generated. Finally, as multi-species approaches remain scarce, there is a need to  
214 define hypotheses at the beginning of such investigations. The integration of  
215 mathematical and statistical models with fish population genetics would be useful for  
216 revealing genotype-phenotype interactions (Ritchie et al., 2015), evolutionary  
217 signatures (Stark et al., 2007), functional DNA elements (Schridder & Kern, 2014), and  
218 spatial dynamics (Guillot et al., 2009).

## 219 *eDNA*

220 The use of environmental DNA, or eDNA, to identify the presence and understand  
221 the distribution of fish has expanded rapidly in the last decade. eDNA is a  
222 polydisperse mixture (Turner et al., 2014; Wilcox et al., 2015) of various biological

223 material ranging from entire cellular fragments to extracellular DNA, which is isolated  
224 from environmental samples such as water or sediment. Such techniques are used  
225 for species identification and food security purposes. Universal primers that target  
226 mitochondrial DNA can be applied for identifying species presence (Yamamoto et al.,  
227 2016) or to gain information about species natural history (e.g. food web construction,  
228 Sousa et al. (2016)).

229 An important component of this work is validating the results from eDNA surveys with  
230 traditional fish survey methods. In both freshwater and marine environments, eDNA  
231 has compared favourably to traditional fish survey methods (Eichmiller et al., 2014;  
232 Sigsgaard et al., 2017). However, eDNA was found to be less effective compared to  
233 experienced snorkel surveys (Ulibarri et al., 2017). This underpins the importance of  
234 validation with traditional techniques, especially in spatially heterogeneous and  
235 complex aquatic environments (Shogren et al., 2017).

236 The development of effective PCR primers is central to the successful application of  
237 eDNA (e.g. Freeland, 2016; MacDonald & Sarre, 2017). As a result, a vast range of  
238 primer sets are available for fishes (e.g. Clusa et al., 2017; Gargan et al., 2017; Doi  
239 et al., 2015). Metabarcoding primers, that simultaneously amplify eDNA from many  
240 fish species, have also been developed for monitoring entire fish communities (Miya  
241 et al., 2015; Valentini et al., 2016).

242 Beyond inferring if a fish species is present in the sampled location, researchers have  
243 begun to investigate if eDNA can provide further information regarding fish  
244 populations. The use of eDNA to infer population level variation has been  
245 demonstrated (Sigsgaard et al., 2016; Uchii et al., 2015), but is still in its infancy.

246 Similarly, several attempts have been made to link eDNA concentration and fish  
247 biomass (Lacoursière-Rousse et al., 2016; Thomsen et al., 2016; Yamamoto et al.,  
248 2016), producing encouraging results. Further development is required to use this  
249 tool to more accurately estimate fish abundance and biomass.

250 However, for techniques utilising eDNA to be optimised, preexisting molecular  
251 information needs to be accessible. A number of publicly available databases (e.g.  
252 NCBI Genbank and BOLD - boldsystems.org) hold a vast array of molecular data but  
253 there is still a need for further mitochondrial genome sequencing to allow optimal  
254 usage of molecular identification techniques (but see Deiner et al., 2017).

## 255 *Microbiomes*

256 Analysis of a microbiome can provide novel insights into the health and biology of fish  
257 populations. Traditional culture-dependent tools used to map the commensal  
258 microbiota community in teleost fish are often time-consuming, expensive and  
259 subjected to bias as only 0.1-10% of bacteria can be cultured *in vitro* (Amann et al.,  
260 1995; Austin, 2006). More recently, rapid culture-independent tools such as 16S  
261 rRNA targeted sequencing have been utilised to provide detailed profiles of the  
262 structure and diversity of the microbiota residing on the mucosal surface of fish  
263 (Ghanbari et al., 2015).

264 The gut microbiome composition has also become an important biomarker for  
265 understanding the influence of stress in fish (Llewellyn et al., 2014), as numerous  
266 stressful stimuli have been shown to alter the microbiome composition (Xia et al.,  
267 2014; Gaulke et al., 2016). The gut microbiome composition can provide insights into  
268 the ecology and physiology of fish in a range of areas such as ecological speciation

269 (Sevellec et al., 2014), the biology of migratory fish (Llewellyn et al., 2016), trophic  
270 interactions within ecosystems (Ingerslev et al., 2014) and adaptation to extreme  
271 environments (Song et al., 2016).

272 There are a number of challenges currently facing fish microbiome research. At  
273 present, the majority of data regarding the microbiome composition in wild teleost fish  
274 originates from laboratory models (Tarnecki et al., 2017). More studies are required  
275 to see if captive-reared animals provide a reliable analogue for wild populations.  
276 Standardised protocols for collecting and generating microbiome data are also  
277 lacking, which could restrict progress as several processes have the potential to  
278 introduce differential bias in microbiota profiles (e.g. Salipante et al., 2014; Hart et al.,  
279 2015; Lyons et al., 2017). Adopting a framework of robust, quality-controlled  
280 protocols (e.g. similar to human microbiome research Methé et al., (2012)) would be  
281 of great benefit. In addition, there is currently a lack of non-invasive protocols for  
282 conducting longitudinal or repeated sampling of the gut microbial community in  
283 individual fish over time. The application of rectal swabs (Budding et al., 2014) for  
284 sampling the vent of fish could provide a non-invasive strategy for collecting  
285 microbiome data from individuals over time. Finally, time-series data could also  
286 enhance our knowledge in terms of the functional aspects of host lifecycles and the  
287 stability and resilience of microbiota (Goodrich et al., 2014).

## 288 **Survey Tools**

### 289 *Field-based surveys*

290 Fish population assessments are conducted using a wide range of techniques; the  
291 advantages, limitations, personnel requirements and health and safety

292 considerations of each are presented in Table 1. It is encouraging to note that even  
293 well-established methods such as hydro-acoustics are continually being improved,  
294 while emerging tools such as eDNA (see above) are beginning to be included in  
295 routine monitoring. We suggest that integrating methods and data series are key  
296 priorities for future research in this field.

297 In large and complex habitats it is often the case that a suite of survey methodologies  
298 has to be employed to sample different times, places and fish species effectively.  
299 Indeed, an advantage of field-based surveys is the ability to generate information  
300 from both fishery-independent (Nash et al., 2016) and fishery-dependent (Shin et al.,  
301 2010) data. However, this need, and the availability of a diversity of methodologies,  
302 can make the task of assessment in these habitats even more costly; issues also  
303 remain over how to use often disparate data types to develop a sound understanding  
304 of a fishery. Integrating methods represents a key means of improving data resolution  
305 from such field surveys. For instance, methods such as eDNA and hydro-acoustic  
306 sampling provide comparatively fast and non-invasive estimates of fish community  
307 structure and biomass. However, to obtain a thorough understanding of fish  
308 populations, this information must be combined with fish age, size and health data  
309 obtained via destructive sampling (e.g. gill netting). As yet, there are no structured,  
310 universally agreed guidelines on which methods should be integrated to obtain a  
311 thorough assessment of population dynamics from a specific habitat type.

312 Fish survey methodologies are typically determined at a national level, making  
313 international comparisons of data extremely challenging. In recent years,  
314 standardised protocols initiated through the EU Water Framework Directive have  
315 facilitated Europe-wide assessments of fish community structure. Such international

316 standardisation is essential when assessing the impact of anthropogenic effects on  
317 fish (see Gordon et al., this issue), and we recommend that efforts are made to make  
318 national datasets available using standardised metadata and biodiversity information,  
319 ideally via open sharing platforms (e.g. <http://www.freshwaterplatform.eu/>).

### 320 *Historical records*

321 Historical records (e.g. catch records) can also be useful in helping to extrapolate  
322 population data back into the recent past. Libraries and historical societies often hold  
323 picture archives and these images can in some instances be used as a form of  
324 historical survey data to provide information on past community composition and size  
325 distributions (McClenachan, 2009). Historical records of catch data are typically held  
326 by government agencies or can be found in local archives (e.g. angling club logs)  
327 and corporate records. Such data have been used successfully to reconstruct fish  
328 populations back to the late 1800s (Thurstan & Roberts, 2010; Thurstan et al., 2010).  
329 Catch reconstruction approaches can also provide useful insights into fishery trends  
330 that may not be apparent from Food and Agriculture Organization (FAO) reported  
331 data alone (Smith & Zeller, 2015; Zeller et al., 2015). Although limited to the  
332 information that is still available and subject to the often-unidentifiable biases of the  
333 individuals who originally recorded the data, such data can provide a unique way to  
334 extrapolate population data back in time.

### 335 **Statistical and modelling tools**

336 *Bayesian methods* - Reliable estimates of demographic parameters (e.g. abundance,  
337 survival, growth rates and fecundity) and an understanding of the processes that  
338 regulate these parameters are fundamental for sustainable management of fish

339 populations. However, to understand the ecological processes and to truly inform  
340 policy, researchers must use multiple data sources, provide links between  
341 management actions and population responses and also estimate uncertainty as a  
342 prerequisite to making forecasts that provide useful information. Bayesian methods in  
343 ecology and conservation biology are now increasingly being used to explore these  
344 links, for example, in stable isotope analyses (see below). Indeed, the Bayesian  
345 framework provides an intuitive method for estimating parameters, expressing  
346 uncertainty in these estimates and allows for the incorporation of as much or as little  
347 existing data or prior knowledge that is available (Ellison, 2004). However, to develop  
348 the use of this specific framework in fish ecology and management, there is a need to  
349 educate and train fish biologists in the use of Bayesian principles and methods.

350 *Individual-based models (IBMs)* are process-based mechanistic computer models  
351 that simulate emergent properties of fish biology, behaviour, traits or group  
352 characteristics, based on simple heuristic functions, and their use has grown  
353 exponentially (e.g. DeAngelis & Mooij, 2015) as computational power has increased  
354 (DeAngelis & Grimm, 2014). Several separate individual-based models were  
355 presented at the 50th Symposium of the FSBI, and, with continued increases in  
356 computational power, IBMs look set to offer powerful new avenues for population  
357 research (DeAngelis & Grimm, 2014) in computationally challenging multifactor  
358 systems such as fish ecotoxicology (e.g. Mintram et al., 2017). Additionally, a variety  
359 of tools now exist which provide for the easier creation of new models, such as  
360 various R packages (see: <http://derekogle.com/fishR/packages>) and programmable  
361 environments (e.g. NetLogo; <https://ccl.northwestern.edu/netlogo>). However, programs  
362 such as R are sometimes not intuitive to new users, and so additional training for  
363 fisheries scientists and collaborations between scientists from different computational

364 and statistical backgrounds would be advantageous. For more robust future  
365 application of IBMs within fisheries science, there is a need for more assessment of  
366 the relative strengths and weaknesses (and potential availability and future  
367 development) of the different models.

368 Integration with environmental data is a pertinent issue when modelling and is  
369 becoming easier through developments in GIS and other programming environments  
370 (such as R), which now include procedures and libraries for use in ecological work.  
371 One example is the use of food web models that integrate environmental data (e.g.  
372 Christensen & Walters, 2004) and coral reef ecosystem modelling methods (e.g.  
373 Rogers et al., 2014; Weijerman et al., 2015). A hindrance to the integration of  
374 environmental data into fisheries science is that it can be difficult to find and access  
375 data sources, although availability and accessibility of such data is improving (e.g.  
376 worldclim.org). The existence of a central node or hub with paths to these data  
377 sources would be useful.

## 378 **Tissue analysis**

### 379 *Stable isotope ecology*

380 Stable isotopes are now routinely used to quantify the trophic ecology (Boecklen et  
381 al., 2011) and migration history (Trueman et al., 2012) of fishes, or to identify  
382 community level patterns in food web structure and resource use (Layman et al.,  
383 2012). Although the technique is still in its relative infancy, stable isotope ecology has  
384 advanced much in recent decades. Below we outline four areas of rapid development  
385 with potential to enhance the applicability of this tool to studies of fish biology.

386 *Biochemical mechanism:* The relationship between the isotopic composition of a  
387 consumer's tissues and that of its prey is fundamental to all applications of stable  
388 isotopes in ecology. However, while general principles are clear (i.e. faster reaction  
389 rates and preferential incorporation of light isotopes into excretory metabolites a  
390 process termed trophic fractionation (DeNiro & Epstein, 1977)), the precise  
391 mechanisms leading to fractionation and, particularly, the extent of isotopic  
392 fractionation expected under differing physiological conditions cannot currently be  
393 predicted, primarily due to the complexity of amino acid biochemistry. Uncertainties  
394 associated with the isotopic expression of tissue composition, and relative rates of  
395 tissue growth and regeneration further complicate the interpretation of stable isotope  
396 values in ecology. However, recent information gained from compound-specific  
397 isotope analysis (i.e. assessing isotopic compositions of single amino acids) is  
398 beginning to shed light on the fractionation process (McMahon & McCarthy, 2016).

399 *Population-level data:* The distribution of isotopic compositions of individuals within a  
400 population (often termed the 'isotopic niche', Newsome et al., 2007) has been  
401 proposed as a powerful comparative measure of population-level ecological  
402 characters. However, in addition to individual variability in consumers, the distribution  
403 of isotopic compositions in a population is influenced by spatial and temporal  
404 variations in the isotopic composition of primary production, temporal variability within  
405 trophic linkages and differential rates of growth and isotopic assimilation. Very few  
406 studies have attempted to combine ecological and food web theory with isotope  
407 systematics to explore the sensitivity of community isotopic metrics to changes in  
408 food web structure and function.

409 *IsoBank*: To date, applications of stable isotopes to fish biology have predominantly  
410 focussed on analyses of specific populations or communities. The absence of a  
411 centralised, open-access repository for stable isotope data restricts the opportunity  
412 for syntheses or meta-analyses of stable isotope data (Pauli et al., 2017). Recent  
413 efforts to address this have found broad support from the stable isotope research  
414 community (Pauli et al., 2017) and would be especially beneficial to fish biologists  
415 due to the large amount of fish isotope data currently available. Defining an ontology  
416 of stable isotope metadata, information required to describe and interpret isotope  
417 data, for fish biologists is an immediate requirement in this regard.

418 *Marine isoscapes*: The stable isotope ratios of a consumer's tissue encode the  
419 resources (water, air, prey etc.) it was using when that tissue was formed. As such,  
420 provided one has access to a suite of isotopic baseline measurements (e.g. water,  
421 plants and primary consumers), it is possible to trace an organisms route through  
422 space and time up to the point of capture (Trueman et al., 2012). Creation of a  
423 practically useful isoscape requires relatively dense sampling of a reference  
424 organism across space (and potentially time). Bulk stable isotope analyses are now  
425 routine, commonly available globally, and relatively cheap, and regional marine  
426 isoscape models are being developed at a rapid rate (MacKenzie et al., 2014; Kurle  
427 & McWhorter, 2017). In the open ocean, sample-based isoscapes are difficult to  
428 develop, but progress is being made in isotope-enabled global biogeochemical  
429 models (Magozzi et al., 2017), offering temporal and spatial models of expected  
430 isotopic variability at global scales. Improving the precision, accuracy and availability  
431 of these baseline measurements will increase the robustness and precision of  
432 isotope based estimates animal position.

433

434 *Archaeological material*

435 Archaeological material can allow an otherwise impossible snapshot into past  
436 populations. Traditional morphological approaches can provide age distributions and  
437 species ranges, and with the explosion in the past 20 years of biomolecular  
438 archaeology, many of the techniques used to explore modern populations can now  
439 be used to look into the past. From ancient DNA to proteomics, and isotopes to lipids,  
440 a wide range of biomolecules have been recovered and explored from archaeological  
441 material (Orton, 2016). For example, compound-specific isotope analysis has the  
442 potential to track trophic level changes through time (McClelland & Montoya, 2002;  
443 Naito et al., 2016). Population genetics of extinct populations have been successfully  
444 explored in terrestrial animals (Chang & Shapiro, 2016; Murray et al., 2017) and  
445 these same techniques can be used on fish bones to reconstruct past genetics  
446 (Iwamoto et al., 2012; Ólafsdóttir et al., 2014). Ideally these data will be used to  
447 understand environmental and anthropogenic effects on fish populations and how  
448 modern fish populations might respond to climate change and fishing pressures.

449 A major barrier to the use of archaeological fish material is the fact that less than  
450 10% of fish bones are identified to species (Wheeler & Jones, 1989; Gobalet, 2001)  
451 and much of what is identified is buried in the 'grey literature' of archaeological  
452 reports that are often not digitised and printed in small quantities (Linden & Webley,  
453 2012). This makes the material relevant to an ecological question very difficult to find.  
454 Archaeologists are working towards ways to improve the amount of bones identified  
455 by better reference collections and education on fish bones (National

456 Zooarchaeological Reference Resource, Nottingham's Archaeological Fish  
457 Resource, Vertebr@UWF) and on creating searchable databases of archaeological  
458 material (Callou, 2009; Kansa, 2010). In addition, new ZooMS (Zooarchaeology by  
459 Mass Spectrometry) techniques are being explored to quickly identify even small  
460 bones and scales to species using peptide mass fingerprinting (Richter et al., 2011)  
461 which will allow even more material to be identified in a useful way for those working  
462 on understanding fish populations. In the near future, it should be possible for  
463 modern fish biologists, in conjunction with archaeologists, to ask direct questions of  
464 past populations (Van Neer & Ervynck, 2010).

#### 465 **General topics identified as applicable across all themes**

##### 466 *Management of data: integration, calibration and standardisation*

467 Progression of an integrated management framework for data classification,  
468 characterisation, storage and accessibility would be a valuable resource for fish and  
469 fisheries biologists. FishBase, which at the time of writing contains information  
470 regarding 33,600 fishes, involving 2290 collaborators, and receives over 600,000  
471 visits per month, is an example of the potential for such a resource (see:  
472 fishbase.org; Froese & Pauly, 2017). A single database for all types of fish data (for  
473 example, DNA, tagging, isotopes, diet) is probably unworkable, but the advent of  
474 application programming interfaces (API) and analytical software which allows  
475 automated querying across multiple databases represents an unprecedented  
476 opportunity to access a wealth of global data. Indeed, we suggest that more data  
477 (such as those discussed here) could be integrated into FishBase. However, such

478 resources require significant funding and long-term commitment from governments  
479 and trans-national organisations, e.g. NASCO.

480 *Public engagement, education and outreach*

481 Scientific engagement with the public is essential to effect meaningful societal  
482 change or to ensure a wider consensus is made around new discoveries or ethical  
483 considerations. Additionally, however, the power of the public as a “tool” in science is  
484 also being increasingly recognised. ‘Crowdfunding’, whereby a scientist requests  
485 small amounts of money from a large number of interested individuals to successfully  
486 launch a project, potentially provides a powerful new way to raise funds, overcoming  
487 some of the difficulties of raising money from traditional grant bodies, especially for  
488 early career researchers or those in developing countries (Wheat et al., 2013).

489 In addition to funding science, the public can also actively engage in the process of  
490 research directly through citizen science projects. Whilst research conducted by non-  
491 professionals is certainly not a new concept, the numbers of projects involving citizen  
492 scientists are growing, especially in the fields of environmental science and ecology  
493 (Silverton, 2009). Through catch records of amateur anglers and commercial net  
494 fishery data extending back many years, research into fish and fisheries is uniquely  
495 placed to benefit from citizen science projects (Stuart-Smith et al., 2013), which have  
496 effectively spanned generations of contributors. Similarly, REEF (reef.org) has been  
497 collecting reef fish diversity and abundance data from trained volunteer divers for 27  
498 years, and the data have been successfully leveraged in hundreds of publications  
499 (e.g. Stallings, 2009; Serafy et al., 2015). Citizen science can also help achieve  
500 important social outcomes, e.g. in establishing sustainable fisheries and marine

501 protected areas, MPAs (Bonney et al., 2014). And, as with crowdfunding, the best  
502 examples of citizen science typically encourage deeper engagement with the public,  
503 and offer a pathway to the democratisation of science.

#### 504 *Fisheries policy and governance*

505 Conserving critical habitats is central to the sustainable management of fish species  
506 and populations. Marine Protected Areas (MPAs), networks of MPAs and Marine  
507 Conservation Zones (MCZs) are widely accepted management tools for fish and  
508 other marine organisms that have been established in many countries (Harborne et  
509 al., 2008; OSPAR, 2013). However, the design of MPA networks could benefit greatly  
510 from the integration of traditional survey data, along with modelling and connectivity  
511 data (Botsford et al., 2009; Grüss et al., 2014). From a social science perspective,  
512 there is a need to better understand public perceptions of marine-related  
513 conservation issues, e.g. fishery regulations, MPAs and MCZs, and to incorporate  
514 these data into fisheries policy and governance frameworks. For example, there is  
515 high public support for MPAs, with surveys showing that people desire around 40% of  
516 the UK's marine waters to be protected (Hawkins et al., 2016). But, while the public  
517 appears to realise that in reality levels of coverage are well below 40%, there is still a  
518 substantial disconnect between perceived coverage of highly protected UK MPAs  
519 (11%) and actual MPA coverage (<0.1%); ultimately, this means that people believe  
520 the UK oceans receive a higher level of conservation than in reality they do (Hawkins  
521 et al., 2016). Developing and implementing effective policies for fisheries  
522 management remains challenging because of the complexities of fisheries and the  
523 socio-political landscape under which they typically operate (Jentoft & Chuenpagdee,  
524 2009). However, the establishment of guidelines or frameworks for fisheries policy

525 and governance (e.g. FAO Voluntary Guidelines for Securing Sustainable Small-  
526 Scale Fisheries) have the potential to better address these challenges and provide  
527 appropriate implementable solutions.

528

## 529 **Conclusions**

530 Across all five of the research themes identified here, it is clear that innovative and  
531 novel tools are being employed to understand all aspects of the biology of fish  
532 populations. Notwithstanding, the authors call for the continued development of these  
533 new and emerging techniques. In particular, there is a need for better integration of  
534 these methods and resulting data, to inform scientifically sound management and  
535 conservation of fish populations.

536 However, it should be noted that, not infrequently, revolutionary methods have been  
537 pedestalled as providing the ability to offer unprecedented novel answers to long-  
538 standing practical problems. Unfortunately, the danger is that such methods can (by  
539 their novelty and the excitement surrounding them), blinker scientists into posing  
540 questions that showcase the methodology, rather than the biology (for example, the  
541 plethora of papers that emerged in the early 1990s extolling the virtues of the random  
542 amplified polymorphic DNA (RAPD) technique). The potentially reduced power of  
543 using any technique on its own (new or otherwise), in isolation of other apparently  
544 'antiquated' methods can turn out to be unnecessarily restrictive. Every technique  
545 has its limitations, but often the restrictions of one tool can be substantially alleviated  
546 by the inclusion of another approach (e.g. Godwin et al., 2016), the marriage of which  
547 can provide a new angle for researching challenging biological problems. It is

548 important that both traditional and emerging tools remain in the toolbox of fish biology  
549 research.

550 Likewise, when genetic-based assignment became popular, many researchers  
551 naively believed the days of tagging fish were over. It is now realised that due to the  
552 many stochastic drivers of population structure, genetic stock identification-based  
553 methodologies such as genetic assignment, do not always succeed. In such cases,  
554 there remains a significant role for tagging in fisheries research. As tag sizes  
555 decrease, and the deleterious effects of tag insertions on fish also decrease, we can  
556 anticipate that genetics and tagging will both continue to have a role to play. The  
557 importance of the relative roles of each technique will depend on the questions being  
558 addressed, the population structure of the study species, and the scale of the  
559 questions being assessed.

560 A final example, which highlights the importance of applying inter-disciplinary and  
561 complimentary tools for understanding fish populations, was a five-year, multi-  
562 agency, EU-funded project investigating the migration and distribution of Atlantic  
563 salmon (*Salmo salar* L.) in the north-east Atlantic (the SALSEA project; NASCO  
564 2008). The purpose was to understand not just where salmon go, but what they eat,  
565 migration routes to feeding grounds, and which waters and regions they pass  
566 through. The SALSEA project used a combination of genetics (microsatellites), stable  
567 isotope analysis, at-sea trawls, tagging and gut contents analysis to assess the  
568 movements and diet of Atlantic salmon across the north-east Atlantic Ocean. As a  
569 result of applying these combined approaches, salmon post-smolt movements have  
570 been confidently ascertained (Gilbey et al., 2017). Nonetheless, even while this  
571 comprehensive study was being finalised, a similarly broad-ranging study was also  
572 being undertaken using SNPs (Bourret et al., 2013). Arguably, this method offers

573 both the potential for finer levels of stock discrimination and the ability to better  
574 explore patterns among functional loci, which may make microsatellite-based  
575 analysis redundant within a short period of time (though see Narum et al., 2008).

576 Thus, the authors consider the continued development of emerging tools, together  
577 with the use of multiple methodologies and inter-disciplinary approaches, to represent  
578 the best avenues for further improving our understanding of fish populations. We  
579 implore scientists from unrelated fields to collaborate on such projects. The FSBI 50<sup>th</sup>  
580 Anniversary Symposium represented one such event, where fish-focused  
581 researchers across diverse fields, came together to advance the state of fish biology.

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587

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