

Manuscript Draft

Manuscript Number:

Title: Options for producing a warm-water fish in the UK: Limits to "Green Growth"?

Article Type: Special Issue: Sustainable Food

Corresponding Author: Dr David Colin Little, Ph.D

Corresponding Author's Institution:

First Author: David C Little, PhD

Order of Authors: David C Little, PhD; David Colin Little, Ph.D; Francis J Murray, PhD; Ekram M Azim, PhD; William A Leschen, MSc; Kathleen A Boyd, MSc; James A Young, PhD; Andrew Watterson, PhD

Abstract: This paper explores the development of a sustainable production system for tilapia and the research implications involved with ensuring commercial viability of such a system for UK farmers. The tilapia is a warm water fish with firm texture, white flesh and mild taste quite similar to a cod or haddock. Whilst tropical in origin it is thought to be highly suitable for low cost aquaculture in temperate zones with the potential to be a more sustainable source of food with fewer environmental impacts than other substitutes. Drawing on a literature review and findings from technical trials the paper will review and compare two production systems - novel Activated Suspension Technology (AST) and conventional Recirculating Aquaculture Systems (RAS) - considering their feasibility in terms of potential and financial viability for scaling up to commercial production of tilapia and their environmental and sustainability benefits.

The review concludes that AST based only on microbial floc is currently uncompetitive with RAS in a UK context although the approach has benefits that might be incorporated in a new generation of mixed systems. Refinement of such systems needs to occur with potential adopters and could be part of diversification of mixed farms. Such development might further enhance the ethical values of fish produced in small-scale, modular RAS.

Table 1 Water and land productivity of tilapia RAS compared to selected intensive open-water production systems (after Phillips *et al.*, 1991 reported in Timmons *et al* 2002).

		Water productivity kg m ³⁻¹	Production intensity mt ha ⁻¹ yr ⁻¹	Ratio of land and water use to RAS use	
				Water	Land
RAS	Nile tilapia	10	1,340	1	1
Intensive ponds	Nile tilapia	0.05	17.4	200	77
	<i>Paneid</i> shrimp	0.05 - 0.09	4.2-11	110-200	120-320
	Channel catfish	0.2-0.3	3	400-500	446
Raceways	Rainbow trout	0.005	150	2,100	9

Table 2. Comparison of production parameters in experimental RAS and AST grow-out systems for *O. niloticus* (source: Murray et al. 2007)

Parameter	Source			
	Murray <i>et al</i> 2007	Hargreaves 2006	Rackocy 2002	Avnimelech 1999
Production system	RAS	AST	AST	AST
Indoor/ Outdoor	indoor	indoor	outdoor	outdoor
Dietary crude protein %	30.4	32	32	20
Secondary carbohydrate source	na	none	none	cellulose
Solids management (TSS mg l ⁻¹)	< 2	250-1000	898 (100-1960)	no data
Culture unit	2.8m ³ tanks	1.5 m ³ tanks	200 m ³ tank	50m ² earthen pond
Culture days	107	no data	201	30
Mean temperature (°C)	29.4	no data	28.5	-
Mean start weight (g)	19.1	41	73.6	112
Mean end weight (g)	405	134	678	218
Cumulative SGR (%) ^b	2.8	1.27	1.11 ^a	1.31 ^a
Final biomass kg m ⁻³	25.6	9.8	13.7	16.5
Cumulative FCR ^c	1.1	1.83	1.9	2.17
Cumulative PCR ^d	3.1	no data	-	2.18
Survival	98.2	no data	81	94.8
Water productivity kg m ⁻³	8.1	no data	9.7	no data

^a Extrapolated from start and end weights; ^b Specific growth rate; ^c Food Conversion ratio ^d Protein conversion ratio

REVIEW: Options for producing a warm-water fish in the UK: Limits to “Green Growth”?

David C. Little, Francis J. Murray, Ekram Azim, William Leschen, Kathleen Grady, James A. Young and Andrew Watterson.
Institute of Aquaculture, Department of Marketing and Public Health Research Group, University of Stirling, Stirling, FK9 4LA, U.K. (Tel +44 1786 467923; fax: +44 1786-467200; email: dcl1@stir.ac.uk)

This paper explores the development of a sustainable production system for tilapia and the research implications involved with ensuring commercial viability of such a system for UK farmers. The tilapia is a warm water fish with firm texture, white flesh and mild taste quite similar to a cod or haddock. Whilst tropical in origin it is thought to be highly suitable for low cost aquaculture in temperate zones with the potential to be a more sustainable source of food with fewer environmental impacts than other substitutes. Drawing on a literature review and findings from technical trials the paper will review and compare two production systems - novel Activated Suspension Technology (AST) and conventional Recirculating Aquaculture Systems (RAS) - considering their feasibility in terms of potential and financial viability for scaling up to commercial production of tilapia and their environmental and sustainability benefits.

The review concludes that AST based only on microbial floc is currently uncompetitive with RAS in a UK context although the approach has benefits that might be incorporated in a new generation of mixed systems. Refinement of such systems needs to occur with potential adopters and could be part of diversification of mixed farms. Such development might further enhance the ethical values of fish produced in small-scale, modular RAS.

Introduction

Seafood consumption in the UK is on the rise (Seafish, 2006), although in comparison with other countries the amounts consumed are relatively small. In recent years a high media profile has affected UK consumers’ interest in and perceptions of seafood. Some of the many issues making headlines range from the health benefits of including fish in the diet (Britton, 2006), to concerns with the safety of consuming both wild and farmed fish (Foran, Carpenter, Hamilton, Knuth & Schwager, 2005). Declining wild fish stocks (Worm et al., 2006) and the quality of the marine environment (Royal Commission on Environmental Pollution, 2004) are also frequently brought to the public eye, creating a complex picture for the public. The diversity of contradictory messages received by the public instigates confusion (Young, Grady, Little, Watterson & Murray, 2006).

Most of the fish used for human consumption currently comes from wild capture fisheries; however seafood from aquaculture is growing rapidly and is set to account for 50% of the world’s food fish in the near future (FAO, 2007a). The rapid growth in aquaculture production is attributed to declining wild stocks even as these continue to be exploited for use in feed for farmed carnivorous aquatic species as well as for other forms of intensive livestock production. This is a major cause of controversy (White, O'Neill & Tzankova, 2004)

The largely negative view of aquaculture in the UK, as a highly intensive, specialised and vertically integrated business model contrasts with traditional practice elsewhere around the world. In Asia where global aquaculture remains concentrated, low trophic species such as carps and tilapias still dominate farmed production and much of this is based on pond-based semi-intensive or extensive systems. This type of aquatic farming is characterised by the high proportion of feed being produced through natural food webs *in situ* (Azim & Little, 2006). Traditionally aquaculture was one component of mixed farming systems and geared to meet subsistence and local market needs (Beveridge & Little, 2002). But soaring demand and limitations of these systems has fuelled a major scale-up in the world wide production of farmed 'seafood' over the last two decades both to meet local and, increasingly, international markets. The shrimp boom in the mid-1980s-90s based on a limited number of species (mainly *Penaeus* spp.) and more latterly Nile tilapia (*Oreochromis niloticus*) and Asian river catfish (*Pangasius hypophthalmus*) have both spread and intensified, particularly in developing countries where land, water and labour are abundant and cheap.

Tilapias have been heralded as a seafood commodity with major potential (Josupeit, 2005). In contrast to shrimp production the rapid scale-up in tilapia production has attracted little criticism from environmental groups and instead been portrayed as a white fish alternative to species higher up the food chain (Marine Conservation Society, 2006). They are being produced in a wide range of production systems and countries in the Tropics and Sub-tropics unlike the Asian river catfish where significant production is concentrated in one area-the Mekong Delta of Vietnam. It might be argued that these factors increase the relative opportunity for sustained growth of tilapia production, especially as despite the levels of growth, prices have remained relatively firm (Josupeit, 2005).

Global production of tilapias has soared over the last decade (FAO, 2007b) with particularly significant growth in South America for export markets and China for both internal and export markets (Josupeit, 2007a, b). This tropical species that originated in Africa is now the 6th most popular seafood choice in the USA (National Fisheries Institute, 2005) and major aquaculture producers turn to tilapia as a new species to invest in (Josupeit, 2007b). Although major centres of tilapia production are in Asia, South and Central America and Africa, culture has also become established in North America and Europe in the last few years. Tilapia production in the UK has been mainly characterised by high profile failures to date (Bunting & Little, 2005). This review assesses the technological options for tilapia production within insulated agricultural buildings proposed as a potential option for rural diversification (Little, 2006).

Towards greener aquaculture

In light of the contradictory messages conveyed in the media, consumer understanding of the ethical and human health issues surrounding aquaculture is understandably confused; however there is still a strong desire for fresh, traceable fish amongst UK consumers (Young et al, 2006). The natural shoaling behaviour of many fish species make the farming of fish at high density both practical and ethical (>100

101 kg m⁻³) provided that nutritionally balanced diets can be cost effectively delivered
 102 and the quality of the water can be maintained (Ebeling, Timmons & Bisogni, 2006).
 103

104 Most of the fish species raised intensively are top carnivores most dependent on high
 105 quality feeds conventionally based on fishmeal and oils derived from wild fisheries.
 106 These feed ingredients are subject to contamination with persistent organic
 107 compounds and their amplification through the food chain (Worm et al, 2006). The
 108 relative risk of consumption of such farmed fish compared to fish of wild origin and
 109 other food stuffs for different groups is the focus of increasing consumer and
 110 scientific interest (Foran, Good, Carpenter, Hamilton, Knuth & Schwager, 2005;
 111 Ellingsen & Aanondsen, 2006).
 112

113 The environmental costs of feed and water supply to aquaculture are also becoming a
 114 major cause of criticism (Naylor et al., 2000); particularly for carnivorous species but
 115 intensification of low tropic species such as tilapias and carps is also utilising
 116 increasing amounts of such feeds. So called 'flow through' or 'open' intensive
 117 systems, in which there is little or no water re-use, can be highly polluting on
 118 receiving waters partly because the cost effective removal of dilute soluble nutrients is
 119 problematic. Open systems includes raceways and cages that produce most of the
 120 tilapias traded internationally (Coward & Little, 2001).
 121

122 Rapid global growth in farming fish and shrimp has occurred in tandem with strong
 123 commercial and environmental incentives to reduce the costs of feed and the impact
 124 of effluents respectively. Any review of the short history of aquaculture will illustrate
 125 that intensification is based on increasing the density of stocked animals stimulating
 126 increased use of both water exchange (to maintain water quality) and higher protein
 127 feeds. High water exchange aggravates nutrient loss and restricts opportunities for
 128 recycling these expensive inputs; it has also been linked to poor biosecurity and
 129 spread of disease in shrimp culture. This has caused a paradigm shift in recent years
 130 towards use of lower protein feeds in low exchange, green water systems, initially in
 131 shrimp production (McIntosh, 2000), but increasingly for other species. This
 132 approach appears to be particularly attractive for systems based on low trophic species
 133 such as the tilapias.
 134

135 Recirculating aquaculture systems (RAS) are increasingly common land based
 136 systems in temperate countries in which water is reused after removal of waste
 137 nutrients and heat may be cost effectively retained in the system. The mechanisms for
 138 removal of suspended and dissolved wastes to reduce solids and nitrogenous
 139 compounds hazardous to fish are key parameters of RAS. There is no requirement for
 140 continual discharge of effluents into the environment, as is the case with the majority
 141 of conventional flow-through aquaculture systems. The 'price' to pay is the external
 142 energy cost of moving water through an appropriate water treatment system,
 143 temperature control, provision of adequate dissolved oxygen and need for complete
 144 balanced nutrition. The commercial culture of tilapia in RAS is now established in
 145 North America and parts of Europe as specialised enterprises targeting high value
 146 markets. Such operations have been either based on integration with waste heat or as
 147 stand-alone enterprises (Melard & Philippart, 1981; Bunting & Little, 2005). This
 148 factor together with the fact that they can produce food locally with few effluents
 149 suggests they meet some of the criteria of 'green' food production systems. The recent
 150 history of limited RAS development in the UK suggests that production technologies

and markets are undeveloped; it is however established practice for value-added aquaculture such as accelerated production of juveniles for on-growing in open systems or ornamental production. Elsewhere in Europe they have become more established for catfish and eel production supplying diverse ethnic and cultural markets (Eding & Kamstra, 2002). More limited access to water for UK and European aquaculture and growing regulation on effluents is likely to increase the attraction of RAS. Rapid production cycles for warmwater fish are also attractive. Tilapias can reach marketable size in as little as six months while 18-24 months is the norm for UK farmed rainbow trout (*Oncorhynchus mykiss*) or Atlantic salmon (*Salmo salar*). The extent of potential water and land productivity gains for tilapia cultured in RAS compared with other intensive production systems for warm-water and temperate species are highlighted in Table 1.

Why tilapia, why intensive??

Tilapias have many characteristics amenable to farming, such as its fast growth under a range of conditions, resilience against disease and a flavour and texture comparable with valuable marine fish (Beveridge & McAndrew, 2000). An ability to feed low in the food chain in principle means that production costs can be low but also, importantly, the fish can be marketed to appeal to increasingly informed consumers on environmental and broader ethical grounds. The trends towards more intensive practices by most commercial tilapia producers threatens these potential core advantages but is a response to current commercial realities.

An ability to feed low in the food chain is matched by a high responsiveness to intensification such that tilapias perform well in intensive systems based on complete, but relatively low protein diets. Their tolerance of high densities (lower densities in fact often trigger aggressive territorial behaviour) has meant a rapid uptake of more intensive operations including more use of higher quality supplementary feeds in semi-intensive ponds (Edwards, Yakupitiyage & Lin, 2000) or complete, formulated feeds in intensive systems. Typically only 20-25% of fed protein is retained in the fish raised in intensive systems (Avnimelech, 2006) the balance becoming pollutants that must be removed. In principle if these waste nutrients could be retained in the system they become substrate for protein-rich bacteria that are re-ingested and utilised by the tilapia. Such nutrient recovery *in situ* occurs in conventional ponds but can be operated at a higher level of intensity through use of aeration to maintain microbial floc in suspension. These activated suspension ponds or technology (ASP, AST) have been advocated for both tilapias and shrimp (Avnimelech, Kochva & Diab, 1994; McIntosh, 2000). The nutritional value of such microbial floc to aquatic animals is dependent on several factors: food preference, ability to both ingest and digest it but also the density of the suspended particles (Hargreaves, 2006). Tilapias being both capable of filter feeding and detritivory are ideal candidates for such systems (Dempster, Baird & Beveridge, 1995; Azim, Verdegem, Mantingh, van Dam & Beveridge, 2003).

Potentially the relative operational simplicity of AST can be combined with a production intensity that is economically viable in the context of a diversification option for mixed farms in the UK. Moreover the 'green' characteristics of the approach could be favourable especially as the market for premium ethical food of all

types has developed rapidly but is under-supplied by local producers. The theoretical basis for the AST is now considered before its application in a UK context is assessed.

From feeding to floc

Intensification of aquaculture systems imposes two major technical challenges-the maintenance of dissolved oxygen and the removal of inorganic nitrogenous products. The latter is critical within intensive aquaculture systems as even low levels of unionized ammonia in water are toxic to most cultured species (Timmons, Ebeling, Wheaton, Summerfelt & Vinci, 2002). Oxygen levels typically become the limiting production factor in optimised culture-systems with adequate ammonia/nitrite treatment capacity.

There are three principle nitrogen pathways to remove hazardous N species in aquaculture (1) photo-autotrophic removal by algae, (2) immobilisation by heterotrophic bacteria as proteinacious microbial biomass and (3) chemo-autotrophic oxidation to nitrate by 'nitrifying' bacteria (Ebeling et al., 2006). The relative importance of each varies with system type and production intensity. Hargreaves (2006) distinguishes between 'photosynthetic growth' (PSG) and 'mixed suspended growth' systems (MSG) based on the degree to which water quality is maintained by photosynthetic and bacterial processes. Suspended particulates formed by heterotrophic bacteria also provide efficient substrates for nitrifying bacteria in bio-floc systems. These can be visualised as 'green' and 'brown' water systems.

Suspended-growth systems are further differentiated from 'attached growth' systems as the waste assimilation, recycling and food production occur within the culture unit as opposed to external bio-filters. Most aquaculture occurs in earthen ponds; which can be considered as PSG systems. Conversely, most RAS rely primarily on chemo-autotrophic bacteria attached as aerobic bio-films on filter media. These examples reflect opposing management goals; in attached systems the aim is to remove nitrogen from the system. In suspended systems the aim is to conserve and recycle nitrogen as useful microbial biomass. Suspended growth systems have also been referred to by a range of terms based on biological or containment characteristics: activated suspension ponds (ASP) activated suspension technology (AST), bio-floc technology (BFT), organic detrital algae soup (ODAS) etc.

Intensification of any suspended growth system requires oxygenation and good water mixing to increase the rate of ammonia immobilisation, both of which can be achieved simultaneously through vigorous aeration. Phytoplankton-rich systems will also benefit from *in-situ* oxygen generation, but with intensification they will ultimately become light limited through self shading. Thus sustained aeration and mixing are essential requirements for intensification of both green and brown water systems.

Although few cross-references exist in the literature these processes are also the basis of the 'activated-sludge' sewage treatment process (Ganczarczyk, 1983; Thiel, 2002). The main difference is that bio-floc accumulations in sewage treatment systems are periodically settled and voided in a continuous or semi-continuous process. In closed-AST the goal is to conserve bio-floc as a food source through internal nutrient

recycling. This mode of operation has two further beneficial features. Theoretically, water exchange rates can be reduced compared to conventional RAS, which are themselves conservative consumers of water (Table 1). Secondly, accumulation of waste inorganic nitrogen compounds; unionised ammonia and nitrite (NH_3 and NO_2) will result in growth inhibition or mortality of fish. *In-situ* heterotrophic ammonia and nitrite assimilation therefore also conserves water quality in this vital respect.

These attributes provided the impetus behind two major trends in the development and application of microbial bio-floc systems in aquaculture. The first has its origins in attempts to optimise natural feed production in semi-intensive ponds through various types of bio-manipulation. The second has its basis in the 'zero-water exchange' and water quality remediation possibilities of AST in contexts where water conservation is paramount. This driver had two threads. Researchers in Israel assessed AST as a potential means of simultaneously intensifying yields and water productivity in arid environments. Elsewhere, the same AST features, offered a means of addressing bio-security and environmental concerns associated with shrimp production. The development of intensive 'zero exchange' shrimp systems provided a highly effective means of disease and effluent management (Burford, Thompson, McIntosh, Bauman & Pearson, 2004; Hari, Kurup, Varghese, Schrama, & Verdegem, 2005; Lemonnier & Faninoz, 2006; Samocha et al., 2007) with feed optimisation as a secondary benefit (Burford, Thompson, McIntosh, Bauman & Pearson, 2003; Wasielesky, Atwood, Stokes & Browdy, 2006). The concurrent evolution of these two drivers is considered below.

The limits of natural productivity in ponds were initially explored using input: output work (e.g. Schroeder, 1978) based on the premise that light-limited primary productivity of conventional shallow ponds in the Tropics of $30\text{kg ha}^{-1} \text{d}^{-1}$ could be further enhanced by optimising heterotrophic productivity through addition of carbon rich substrates. Initially, this approach assumed that photo-autotrophic and heterotrophic feed pathways were partitioned and emphasised the role of heterotrophic pathways in achieving further yield gains. However, the interdependence of these pathways and the mechanisms by which fish such as tilapia could filter feed or harvest micro-organisms from the water column soon became apparent (Colman & Edwards, 1987; Avnimelech, Mokady & Schroeder, 1989).

Concurrent work carried out in Israel on more intensive systems suggested that sorghum and other energy-rich grains could be used cost effectively as supplements to natural food-especially micro-algae rather than more protein-rich feeds (Hepher, 1988). Yields in these intensive water-limited systems, were constrained by water quality limits stimulating further work aimed at enhancing AST function. Theoretically, optimising ratios of C:N will enhance conversion of toxic inorganic-nitrogen compounds to microbial biomass available as food for fish or shrimp while further improving water quality. Goldman et al. (1987) elucidated the fundamental nutrient balance principles underlying growth efficiency of marine bacteria. They found C:N ratios $>10:1$ were optimal for optimising bio-floc production while minimising ammonia regeneration. Many investigators (Avnimelech et al, 1989, 1994, 1999, Hari et al., 2004, Burford et al, 2004) then applied this principle as an approach to optimising nutrient inputs and recycling within intensive bio-floc aquaculture systems. The use of a carbohydrate source in addition to conventional feeds or use of

feeds with lower protein content was advocated on this basis for systems in which bio-floc was aerated and retained in the system (Avnimelech, 1999).

The AST concept of further intensification of natural food production and use *in situ* has developed from this practice and theory for species such as tilapias and shrimp that are capable of utilising microbial floc as a major element of the diet and tolerating water high in suspended solids. Higher intensification rates also involve a move from earthen pond systems to lined-pond systems (shrimp) and tanks (tilapia). Most published accounts of AST however, relate to systems which maintain algal-rich water i.e. green water / PSG systems.

Generally green water systems are known to suffer inconsistent water quality, partly related to algal succession that is difficult to control or influence. Bacteria dominated systems tend to be more consistent (Hargreaves, 2006) but the nature and impacts of succession and change within systems with minimal phytoplankton are unknown. Our understanding of low-plankton systems is informed by the experience of managing partitioned aquaculture systems that alternate between autotrophic and heterotrophic status depending on ambient climate (Hargreaves, 2006). The principle of using compartments of algal rich water to remove ammonia is complicated by mixed success in controlling algal biomass.

The adaptation of these principles to a brown water / MSG system in light-limited conditions in which natural feed was mainly bacterial rather than derived from phytoplankton was the major objective of our research. The relative stability of heterotrophic microbial populations and their independence of light conditions on water quality were considered as positive factors (Avnimelech, 2006). For the sunlight-limited seasonal conditions in the UK the concept of well insulated smaller intensive tank-based systems located inside buildings was developed based on such a 'brown-water' approach. We now consider some of fundamental issues that differentiate AST as researched and promoted to date with their potential for use within the farming sector in the UK.

Tilapia as a farm diversification strategy in the UK

Intensive fish culture in the UK has been the preserve of an entrepreneurial business sector and the attraction of this type of diversification for risk-adverse farmers must be considered (Rosa, Kodithuwakku, Young & Little, 2007). Diversification into a novel product (i.e. tilapia) based on a new technical approach (AST) is likely to further increase risk. The potential benefits of using AST for tilapia production rather than RAS scaled down to meet the investment profiles and potential local market niches available to them need to be established.

The costs and risks of maintaining optimal temperatures for warm-water fish are an initial concern to most potential adopters. The optimal temperature range for tilapia production is 28-32°C, however, energy costs (heating and pumping) are proportionately low (15% total direct costs; Timmons, 2005). In the past, RAS have often been linked to waste heat utilisation from distilleries, power stations, factories etc. Whilst an apparently green and cost-effective approach, over-reliance on third-

350 party waste energy has also contributed to failures. A source of low value heat on-
 351 farm may be a motivation for diversification into warm-water fish culture. Another
 352 incentive is the utilisation of disused or underutilised agricultural buildings although
 353 low cost-purpose built structures such as insulated polytunnels also have potential.
 354

355 Reducing the capital requirement and design complexity is an important advantage for
 356 any production system. In principle AST are simpler to design and manage than RAS;
 357 solids (feed and floc) are kept in suspension and dissolved oxygen levels maintained
 358 through aeration. As the culture unit also acts to treat wastes, there is no requirement
 359 for external biofilter, piping or pumps which results in lower capital costs and
 360 theoretically, more straight forward management. The capital outlay of these
 361 components for an RAS can range typically from 10- 35 % of initial fixed costs. Low
 362 cost, simple AST could also be temporary or moveable structures allowing farmers to
 363 take advantage of seasonal availability of space, resources and marketing
 364 opportunities.
 365

366 A potential incentive for producing tilapia using a microbial floc-based system rather
 367 than conventional RAS is the possibility that local feeds can be used. The overall
 368 reduction in feedstock quality required to raise tilapia in AST is potentially a
 369 substantial saving on production costs over RAS in which feed cost typically make up
 370 from 30-40% of total operating costs depending on the scale of the operation and
 371 other factors (Timmons et al., 2002, Timmons, 2005). Using a feed of lower overall
 372 quality feed i.e. 20% crude protein feed rather than typical formulations (28-32%CP)
 373 could reduce reliance on feed ingredients such as fish meal and soybean meals.
 374 Potentially it could open opportunities for growing or using feed ingredients locally or
 375 on-farm in a similar manner to that practiced for intensive dairy production thus
 376 reducing risk and enhancing familiarity that were important priorities for potential
 377 adopters (Rosa et al., 2007).
 378

379 Over-ambitious production schedules, steep technical learning curves and lack of
 380 prior aquaculture experience have been inter-related causes of recurrent failure in
 381 RAS. Contract farming packages which emphasise potential gains while under-
 382 estimating risk has contributed to spectacular failures in other novel farm
 383 diversification start-ups (e.g. ostrich, and Alpaca farming). Research indicates a
 384 similar threat in the UK tilapia sector. Small-scale modular approaches hold potential
 385 for limiting risks carried by new adopters with no previous aquaculture experience.
 386 These adopters then have the option of scaling up to more economically efficient units
 387 required to supply higher volume/ low margin commodity chains (food processors and
 388 supermarkets), or continuing to produce smaller volumes of fresh product for higher
 389 value niche markets. In the US, innovative tilapia production initially targeted value
 390 added markets but relatively high labour costs undermined their capacity to compete
 391 with imports leading them to target specialist live sales, often to ethnic minorities
 392 (Serfling, 2000). Significant scale-up in production of tilapia and other species such as
 393 *Pangasius* spp. in tropical countries threatens competitiveness of producers in the
 394 commodity sector in the UK.
 395

396 A key research question is; can such a production approach be maintained at
 397 production levels that would be cost effective and attractive to farmers in the UK?
 398 The use of aquatic microbial floc as the basis for tilapia production has been
 399 advocated, but research on intensive indoor/ brown-water production systems is still

required to justify promotion of the AST approach to farmers in temperate climates such as the UK.

Comparing performance of AST and RAS

There is a recent history of research on the operation and efficiency of AST systems, most of which is based on intensively fed, green water systems in ponds or tanks (reviewed by Hargreaves, 2006). Most of the commercial application appears to relate to the relatively much lower-density shrimp production with relatively little published information regarding higher density fish production systems (Avnimelech, 2007). Unfortunately there is a dearth of data for replicated large-scale research systems and most conclusions have been drawn based on either short term small-scale experiments and/or observation of commercial or semi-commercial systems based on variable sized fish (Table 2). Only two trials (Rakocy, Bailey, Thoman & Shultz, 2002, Murray et al., 2007) report on-growing to the minimum harvest size of 400g feasible for markets in the UK. Most reports have emphasised the potential for improved feeding efficiency based on nutrient recycling in AST systems compared to RAS or conventional pellet-fed ponds (Avnimelech et al, 1989; Milstein, Avnimelech, Zoran & Joseph., 2001). Avnimelech (2006) for example cites feed: cost ratios in C:N manipulated pond-AST as being almost double control systems with higher crude dietary protein inclusion. However, meaningful evaluation of the commercial potential of AST compared to RAS also requires knowledge of fish growth rates and system carrying capacities. Unfortunately key parameters that would allow interpretation of growth are often lacking (e.g. water temperature) or inadequately presented. In particular crude daily weight gain rather than specific growth rates are routinely used for comparisons using fish of highly variable stocking and harvest weights.

Even allowing for these limitations the magnitude of difference between growth rates is evident. Under controlled temperatures, stocking and feed conditions with C:N manipulation and solids removal, Murray *et al* (2007) found growth rates in AST were only 68% of those achievable in RAS (achieving an SGR of 2.8 % for fish grown from 19g-405g; both systems fed on 30% CP diets). SGRs fell to 36% of RAS levels in AST fed on 18% CP diets. When one accounts for the slower growth rate of larger-fish, grow-out time to 400g is almost doubled in the fastest growing AST compared to the RAS control (Murray et al., 2007).

Low carrying capacities make the commercial case for intensive AST appear still more marginal. Stocking densities exceeding 100 kg fish m⁻³ are routinely achievable in RAS with oxygenation and densities up to 70-80 kg fish m⁻³ with aeration (Timmons et al., 2002). This compares to reported levels of only 10 - 16.5 kg fish m⁻³ in AST (Table 2). Murray et al. (2007) achieved levels of 28 kg fish m⁻³, but only using complete feeds and solids removal. Clearly the benefits of feed and water use efficiencies reported for AST need to be viewed in the context of growth inhibition and reduced carrying capacity in intensive systems. Both factors have consequences for overall production costs when capital and variable costs for building size, floor area, insulation, labour and heating etc. are considered. The same constraints also eliminate gains in water efficiency; Murray et al. (2007) and Rakocy et al. (2002) measured broadly comparable optimal rates of 7.2 and 9.7 kg m⁻³ achievable in AST compared to typical RAS rates of 8-10 kg m⁻³ (Tables 1 and 2).

The potential for further intensification appears to be fundamentally limited by biological factors which correlate with bio-floc concentration in closed systems. Hargreaves (2007) observed process instability at feeding levels above 200 g m^{-3} equivalent to a stocking density of 10 kg m^{-3} (at a feed rate of $2\% \text{ bw day}^{-1}$). Rakocy et al. (2002) and Murray et al. (2007) observed severe growth inhibition and increased mortality at TSS levels above 850 mg l^{-1} . In practice therefore, there is a requirement for solids removal in AST to maintain a level of suspended solids which will not significantly retard food intake and growth or constrain economically viable stocking densities. This requires some form of external clarifier (Murray et al., 2007; Hargreaves, 2006; Rakocky et al., 2002). However, the variable quality (size, consistency and specific gravity) of microbial floc that occurs over time complicates the design and operational management of such clarifiers in indoor AST systems (Murray et al., 2007).

Operation of AST incorporating solids removal also represents a partial step-back towards RAS-type compartmentalisation with semi-continuous or continuous water re-circulation. Solids settled in external clarifiers could be removed or managed entirely independently for controlled release to the grow-out compartment. Investigators found floc composition varied in closed culture-systems with implications for chronic and acute event-mortalities (Murray et al., 2007; Azim, Little & North., 2007; Rakoky et al, 2002). Compartmentalisation could also provide a means of floc-stabilisation potentially incorporating activated-sludge techniques borrowed from the water and sanitation sector where steady-state operation is a critical feature. One commercial producer in the United States has already moved along this route in a hybrid system; maintaining TSS within $70\text{-}130 \text{ mg l}^{-1}$ and achieving net yields of $60 \text{ kg m}^{-3} \text{ year}^{-1}$ (Serfling, 2000); in other words, resorting to use of bio-floc primarily as a low-cost *in-situ* water treatment process with low water exchange requirements.

The fundamental theoretical benefit of AST; improved feed efficiency can also be challenged. Analysis of feed and crude protein conversion and retention indicate that the amounts of microbial floc in a brown water system utilised as feed over a range of commercial stocking densities in fish offered feeds of a range of quality and presentational form were minimal (Murray et al., 2007). This contrasts markedly with the values published by Avnimelech (1999) based on observations of light-driven AST systems but could reflect differences in interpretation of data. Attempts to manage microbial floc production by manipulating C:N ratio, or floc levels through solid removal are also highly variable in the systems described. Azim *et al* (2007) also reported increased feed conversion efficiency in AST compared to RAS systems in which fish were maintained at low densities and fed similar amounts of feed confirming the utilisation of microbial floc by fish.

There are other important characteristics of tilapia culture in AST that deserve mention. The specific conditions of AST appear to favour beneficial bacteria and reduce disease incidence compared to alternative systems. The absence of disease in AST systems has been related to the probiotic nature of microbial floc (Serfling, 2000; Murray et al 2007; Avnimelech and Bejarano, 2007).

The natural habitat of tilapias are turbid water lakes of Africa but the high levels of suspended solids that characterise AST has raised issues regarding the welfare and taste of the fish produced in such systems. The impacts of high levels of microbial floc in AST systems on the taste and welfare of tilapias has been recently assessed. Off-flavours are related to the absorption and accumulation of natural chemicals or compounds (such as geosmin and MBT) through the gills, skin or gastrointestinal tract of fish (Boyd & Tucker, 1998; Gautier et al., 2002). Contrary to common perception culture systems rich in natural food are not necessarily more likely to produce fish with off-flavour (Serfling, 2000; Eves, Turner, Yakupitiyage, Tongdee, & Ponza, 1995). Bue (2005) conducted organoleptic taste trials on fish raised in both RAS and AST and found no perceived differences among fish direct from tanks or after standard depuration techniques in fresh or saline water (Rungreungwudhikrai, 1995).

Generally high levels of suspended solids are related to poor fish welfare as indicated by poor growth, fusion of gill lamellae (Mettam, 2005) and susceptibility to bacterial or parasite infections (Noble & Summerfelt, 1996). Lower feed intakes and performance withstanding, Vincent (2006) found no indication of gill damage on fish raised over extended periods within AST or RAS systems nor differences in tail erosion, scale loss etc characteristic of poor welfare.

Future research needs

Assessment of the development process towards an intensive system for UK farmers to produce and market an exotic food fish species has identified a number of interesting issues. Prototype RAS systems now require testing with potential producers and this will require an iterative action learning approach whereby insights of the adopters are incorporated. Studies on the nature of entrepreneurship give some insight as to the characteristics of potential adopters and whether diversification was driven by need or opportunism (Rosa et al., 2007).

Clearly there are trade-offs in terms of environmental and broader ethical values of fish produced by RAS and AST. Both systems have very limited effluents which through virtue of their nutrient concentration are useful fertilisers (Watten, & Busch, 1984; McMurtry et al., 1997). Further research to quantify the potential synergisms between water and nutrient use in tank based systems and associated high value horticulture is required. Integration with hydroponics has particular market potential as demand for closed cycle, pesticide-free fruit and vegetables increases.

Although fish produced in AST systems had few overt signs of poor welfare, the lower feed intake, slower individual growth and chronic mortalities observed suggest that RAS provided more consistent and optimal conditions. Further development of mixed systems has been advocated in which culture units are partitioned with algae, microbial floc and/or periphyton (e.g. Avnimelech, 2006; 2007; Azim and Little, 2006; Serfling, 2000). Optimisation of floc levels for commercial applications is a research priority.

The value of microbial floc in terms of preventing fish disease problems warrant scientific investigation. Probiotic approaches are now widespread in the market but

the relative control possible in AST and observations of the high health of fish produced makes further investigation worthwhile. Designs in which the natural feed component can be optimised with respect to nutritional quality and energy efficient ingestion, digestion and assimilation should be prioritised. Development with producers in an action research mode is most likely to result in models which are management efficient and adoptable.

Intensive tilapia production is land efficient (Table 1) and may be located in periurban, rather than rural locations. Benefits include the improved access to a range of consumers, potentially reducing marketing costs. Controlled environments leading to improved predictability of production and expected genetic and feeding gains as has occurred in the broiler industry over the longer term are expected to further improve competitiveness compared with other fish species and substitutes (Timmons, 2005)

The market context for tilapia sales in the UK is dynamic. Consumers are increasingly willing to try new preparations and species of fish (Seafish, 2006b), whether it be for health reasons, indulgence or environmental grounds. Potential for tilapia therefore exists, not only in ethnic markets as a fresh or live alternative to frozen imports, but as a locally available 'green' fish product possibly with eco-credentials (Young, et al, 2006). Tilapia also has potential in the food service sector, where novel, exciting fish products are of interest, particularly if they have amenable aesthetic and preparation qualities (Seafish, 2006a). A locally available, small-scale and high quality tilapia supply would therefore meet industry wide interest in fresh, traceable fish supplies, however, a comparative analysis of the relative competitiveness of tilapias produced locally against imports and substitutes is required.

Acknowledgements

This work has been supported by the Rural Economy Land Use Programme of Research Councils UK under the project 'Warmwater Warm water fish production as a niche production and market diversification strategy for arable farmers with implications for sustainability and public health'. Ekram Azim is supported by a Marie Curie Fellowship of the EU (Contract MIFI-CT-2005-008965). The authors thanks the staff of Nam Sai farm, Prachninburi, Thailand for their support during the project.

References

- Avnimelech, Y. (2006). Bio-filters. The need for a new comprehensive approach. *Aquaculture Engineering*, 34 172-178.
- Avnimelech, Y. (2007). Feeding with microbial flocs by tilapia in minimal discharge bio-flocs technology ponds. *Aquaculture*, 264, 140-147.

Avnimelech, Y. (1999). Carbon/nitrogen ratio as a control element in aquaculture systems, *Aquaculture*, 176, 227-235.

Avnimelech, Y, Mokady, S and Schroeder, G.L. (1989). Circulated ponds as efficient bioreactors for single cell protein production *Israel J. Aquaculture Bamidgeh*, 41, 58-66.

Avnimelech, Y, Kochva, M & Diab, S. (1994) Development of controlled intensive aquaculture systems with a limited water exchange and adjusted carbon to nitrogen ratio. *Israel J. Aquaculture Bamidgeh*, 46, 119-131.

Avnimelech, Y. & Bejarano, I. (2007). Probiotic effects of bio-floc technology: Depression of Streptococci Infection of tilapia
<http://floc.aesweb.org/library/sanantonio2007/avinemelech.pdf>

Azim, M.E., Verdegem, M.C.J., Mantingh, I., van Dam, A.A. & Beveridge, M.C.M. (2003). Ingestion and utilization of periphyton grown on artificial substrates by Nile tilapia *Oreochromis niloticus* L. *Aquaculture Research*, 34, 85-92.

Azim, M. E. & Little, D.C. (2006). Intensifying aquaculture production through new approaches to manipulating natural food. *CAB Reviews: Perspectives in Agriculture, Veterinary Science, Nutrition and Natural Resources*, 1, no.062, 23p **URL** <http://www.cababstractsplus.org/cabreviews/index.asp> **DOI** 10.1079/PAVSNNR20061062

Azim, M.E., Little, D.C., North, B.P., 2007. Growth and welfare of Nile tilapia *Oreochromis niloticus* cultured in indoor tanks using Activated Suspension Technique (AST). World Aquaculture 2007, San Antonio, Texas, USA, 26 February-3 March, 2007

Belton, B. Little, D.C. & Grady, K. *in press* Reassessing Sustainable Aquaculture through the Lens of Tilapia Production in Central Thailand, *Rural Sociology*

Beveridge, M. & McAndrew, B. Eds. (2000). *Tilapias: Biology and Exploitation*. Kluwer Academic Publishers, Dordrecht, Netherlands.

Beveridge, M.C.M. & Little, D.C. (2002). History of aquaculture in traditional societies. p.3-29. In *Ecological Aquaculture* (Ed. B.A. Costa-Pierce) Blackwell Science, Oxford.

Boyd, C.E. & Tucker, C.S. (1998). Pond aquaculture water quality management. *Kluwer academic publishers*. Boston, MA .

Britton, S. (2006). Omega-3 gives health foods something to shout about, *Food Manufacture*, 81, 14.

Bue, C.E. (2006). Evaluation of welfare and organoleptic criteria of Nile Tilapia (*Oreochromis niloticus*) grown in Activated Suspension Technology Systems. Thesis. University of Stirling, Stirling, U.K.

648 Bunting, S.W. & Little, D.C. (2005) The emergence of Urban aquaculture in Europe.
649 Chapter 8 In B.A. Pierce, P. Edwards, D. Baker & A. Desbonnet (Eds). *Urban*
650 *Aquaculture* pp. 119-136. Wallingford, UK. CABI Publishing
651

652 Burford, M.A., Thompson, P.J., McIntosh R.P., Bauman, R.H. & Pearson D.C.
653 (2003) Nutrient and microbial dynamics in high-intensity, zero-exchange shrimp
654 ponds in Belize. *Aquaculture* 219, 393-411.

655 Burford, M.A., Thompson, P.J., McIntosh, R.P., Bauman, R.H. & Pearson, D.C.
656 (2004). The contribution of flocculated material to shrimp (*Litopenaeus vannamei*)
657 nutrition in a high-intensity, zero-exchange system. *Aquaculture*, 232, 525-537.
658

659 Colman, J.A. & Edwards, P. (1987). Feeding pathways and environmental constraints in
660 waste-fed aquaculture: Balance and optimization In D.J.W. Moriarty & R.S.V. Pullin
661 (Eds) *Detritus and Microbial Ecology in Aquaculture* (pp. 240-281) ICLARM
662 Conference proceedings 14, Manila, Philippines, ICLARM.
663

664 Coward, K and Little, D.C. (2001). Culture of the aquatic chicken; present concerns and
665 future prospects. *Biologist* 48 (1).
666

667 Dempster, P., Baird, D. J., & Beveridge, M. (1995). Can fish survive by filter-feeding
668 on microparticles? - Energy balance in tilapia grazing on algal suspensions, *Journal of*
669 *Fish Biology*, 47, 7-17.
670

671 Ebeling, J.M., Timmons, M.B. & Bisogni, J.J. (2006). Engineering analysis of the
672 stoichiometry of photoautotrophic, autotrophic, and heterotrophic removal of
673 ammonia-nitrogen in aquaculture systems. *Aquaculture*, 257, 346-358.
674

675 Eding, E & Kamstra, A. (2002). Netherlands farms tune recirculation systems to
676 production of varied species. *Global Aquaculture Advocate*, 5, 52-55.
677

678 Edwards, P., Yakupitiyage, A & Lin, C.K. (2002). Semi-intensive pond aquaculture.
679 M.C.M Beveridge & B.J. Mc Andrew (Eds) *Tilapias: Biology and exploitation*
680 (pp. 377-403) Dordrecht, Kluwer Academic Publ.
681

682 Ellingsen, H. & Aanonsen, S. A. (2006). Environmental Impacts of Wild Caught
683 Cod and Farmed Salmon - A Comparison with Chicken, *International Journal of Life*
684 *Cycle Analysis*, 11, 60-65.
685

686 Eves, A., Turner, C., Yakupitiyage, A., Tongdee, N. & Ponza, S., (1995). The
687 microbiological and sensory quality of septage-raised Nile Tilapia (*Oreochromis*
688 *niloticus*). *Aquaculture*, 132, 261-272.
689

690 FAO (2007a). *The State of World fisheries and Aquaculture 2006*, Food and
691 Agriculture Organisation of the United Nations
692 <http://www.fao.org/docrep/009/A0699e/A0699e00.htm>
693

694 FAO (2007b). Fisheries Global Information System, *Global Aquaculture Production -*
695 *Tilapia*, Food Agriculture Organisation of the United Nations, www.fao.org
696

697 Foran, J., Carpenter, D., Hamilton, M., Knuth, B., & Schwager, S. J. (2005). Risk-
698 Based Consumption Advice for Farmed Atlantic and Wild Pacific Salmon
699 Contaminated with Dioxins and Dioxin-like Compounds. *Environmental Health*
700 *Perspectives*, 113, 552-556.
701

702 Foran, J., Good, D., Carpenter, D., Hamilton, M., Knuth, B., & Schwager, S. J.
703 (2005). Quantitative Analysis of the Benefits and Risks of Consuming Farmed and
704 Wild Salmon. *Journal of Nutrition* 135,2639-2643.
705

706 Ganczarczyk, J. (1983). Activated sludge process: theory and practice. Pollution
707 engineering and technology series; No. 23. Marcel and Decker Inc. New York.

708 Goldman, J.C., Caron, D.A.& Dennett, M.R., (1987). Regulation of gross growth
709 efficiency and ammonium regeneration in bacteria by substrate C:N ratio. *Limnology*
710 *and Oceanography* 32, 1239-1252.
711

712 Grimm, C.C., Lloyd, S.W. & Zimba, P.V., (2004). Instrumental versus sensory
713 detection of off-flavours in farm-raised channel catfish. *Aquaculture*, 236, 309-319.
714
715

716 Hargreaves, J. A. (2006). Photosynthetic suspended-growth systems in aquaculture
717 *Aquacultural Engineering*, 34, 344-363.
718

719 Hargreaves A& Wong, H. W, (2007). The effect of solids concentration on
720 performance of a bio-floc system for tilapia. Paper presented at Aquaculture 2007.
721 Sustainable Aquaculture session, San Antonio Texas.
722
723

724 Hari, B., Kurup, B.M., Varghese, J.T., Schrama, J.W.& Verdegem, M.C.J. (2006).
725 The effect of carbohydrate addition on water quality and the nitrogen budget in
726 extensive shrimp culture systems. *Aquaculture*, 252, 248-263.
727

728 Hepher, B. (1988). *Nutrition of Pond Fishes.*, Cambridge, Cambridge University
729 Press.

730 Huntingford, F.A., Adams, C., Braithwaite, V.A., Kadri, S., Pottinger, T.G., SadØe,
731 P. & Turnbull, J.F. (2006). Current issues in fish welfare. *Journal of Fish Biology* 68,
732 332-372.
733

734 Josupeit, H. (2005), *World Market of Tilapia*, Globefish, FAO, Rome.
735

736 Josupeit, H. (2007a) Tilapia Market Report-China, March 2007, Globefish, FAO,
737 Rome.
738

739 Josupeit, H. (2007b) Tilapia Market Report, June 2007, Globefish, FAO, Rome.
740

741 Lemmonier, H. Faninoz, S. (2006). Effect of water exchange on effluent and sediment
742 characteristics and on partial nitrogen budget in semi-intensive shrimp ponds in New
743 Caledonia. *Aquaculture Research*, 37, 938-948.

744 Little, D.C. (2006). A new approach to tilapia in the UK-based on activated
745 suspension technology and niche markets. *Aquaculture Today* 28-30 th March 2006
746 Edinburgh

747

748 MacIntosh, R.P. (2000). Changing paradigms in shrimp farming: IV:low protein feeds
749 and feeding strategies. *Global Aquaculture Advocate* 2, 40-47.

750

751

752 Marine Conservation Society (2002). *Good Fish Guide*. www.mcsuk.org.

753

754 McMurtry, M.R., Sanders, D.C., Cure, J.D., Hodson, R.G., Haning, B.C. & St
755 Amand, P.C. (1997). Efficiency of water use of an integrated fish/vegetable co-culture
756 system, *Journal of the World Aquaculture Society* 4, 420-428.

757

758

759 Melard, C. & Philippart, J.C. (1981). Pisciculture intensive du tilapia *Saratherodon*
760 *niloticus* dans less effluents thermiques d'une centrale nucleaire en Belgique. In
761 K.Tiews (Ed) *Proceedings of the World Symposium on Aquaculture in heated*
762 *effluents ad recircualtion systems*. Volume 1 (pp 637-659.) 28-30th May 1980
763 Stavanger, Berlin, Heenemann.

764

765 Mettam, J. (2005). An investigation into the use of gill pathologies in rainbow trout
766 (*Oncorhynchus mykiss*) as a welfare score reflecting water quality. Thesis. University
767 of Stirling., Stirling, U.K.

768

769 Milstein, A., Avnimelech, Y., Zoran, M. and Joseph, D. (2001). Growth performance
770 of hybrid bass and hybdrid tilapia in conventional and active suspension ponds. *The*
771 *Israeli Journal of Aquaculture* 53, 147-157.

772

773

774 Murray, F.J and Little, D.C. (2007). Evaluation of a commercial-scale activated
775 suspension production system for intensive tilapia culture in indoor tanks. In Press

776

777 National Fisheries Institute (2007) *Top 10 US Consumption By Species*
778 www.aboutseafood.com

779

780 Naylor, R.L., Goldburg, R.J., Primavera, J.H., Kautsky, N., Beveridge, M.C.M., Clay,
781 J., Folke, C., Lubchenco, J., Mooney, H.& Troell, M. (2000). Effect of aquaculture on
782 world fish supplies. *Nature*, 405, 1017-24.

783

784 Noble, A.C. & Summerfelt, S.T. (1996). Disease encountered in rainbow trout
785 cultured in recirculating system. *Annual Review of Fish Diseases* 6, 65-92.

786

787 Rakocy, J.E. (2002) An integrated fish and field crop system for arid areas In. B.A.
788 Costa Pierce (Ed) *Ecological Aquaculture:The evolution of the Blue Revolution*
789 (pp.263-285).Oxford, UK. Blackwell Science.

790

791 Rakocy, J.E., Bailey, D.S., Thoman E.S.& Shultz, R.C. (2005). Intensive tank culture
 792 of tilapia with a suspended, bacterial-based, treatment process. University of the
 793 Virgin Islands. <http://ag.arizona.edu/azaqua/ista/ista6/ista6web/pdf/584.pdf> . 13pp
 794

795 Rosa, P., Kodithuwakku, S., Young, J.A. & Little, D. (2007). Opportunity, necessity
 796 & entrepreneurial success: a farming perspective, *Proceedings in 4th ASGE*
 797 *International Entrepreneurship Research Exchange*, February 6-9th, 2007, Brisbane,
 798 Australia.
 799

800 Royal Commission on Environmental Pollution (2004). *Turning The Tide: Addressing*
 801 *the impact of fisheries on the marine environment*. Summary Report.
 802

803 Rungreungwudhikrai, E.O. (1995). Characterization and classification of off-flavour
 804 of Nile tilapia. Thesis AE-95-29. *Asian Institute of Technology*. Bangkok, Thailand.
 805

806 Samocha, T.M., Patnaik, S., Speed, M., Ali, A.M., Burger, J.M., Almeida, R.V.,
 807 Ayub, Z., Harisanto, M., Horowitz, A. & Brock, D.L. (2007). Use of molasses as
 808 carbon source in limited discharge nursery and grow-out systems for *Litopenaeus*
 809 *vannamei* *Aquacultural Engineering*, 36, 184-191.
 810

811 Schroder, G.L. (1978). Autotrophic and heterotrophic production of microorganisms I
 812 intensively manured fish ponds and related fish yields. *Aquaculture*, 14 303-325.
 813

814 Seafish (2006). *Foodservice Market Update August 2006*, Seafish Industry Authority.
 815 www.seafish.org.
 816

817 Serfling, S.A. (2000). Closed-cycle , controlled environment systems: The solar
 818 aquafarms story. *The Advocate*, June 48-51.
 819

820 Thiel D.J. (Ed.) (2002). Activated Sludge, WEF Manual of Practice No. OM-9. Water
 821 and Environment Federation, Alexandria.
 822

823 Timmons, M.B. (2005). Competitive potential for USA urban aquaculture. In B.
 824 Costa-Pierce, A. Desbonnet, P. Edwards and D. Baker (Eds) *Urban Aquaculture*
 825 (pp.137-157) Wallingford, UK/Cambridge, Massachusetts: CABI Publishing.
 826

827 Timmons, M.B., Ebeling, J.M., Wheaton, F.W., Summerfelt, S.T., Vinci, B.J. (2002).
 828 Recirculating Aquaculture Systems, 2nd Edition. Caruga Aqua Ventures, New York,
 829 USA.
 830

831 Verdegem, M.C.J., Bosma, R.H.& Verreth, J.A.J. (2006). Reducing water use for
 832 animal production through aquaculture. *International Journal of Water Resources*
 833 *Development* 22(1), 101-113.
 834

835 Vincent, Y.R. (2006). Use of gill condition to assess welfare of tilapia raised in two
 836 intensive production systems. Thesis. University of Stirling, Stirling, U.K.
 837

838 Wasielesky, W., Atwood, H., Stokes, A., Browdy, C. (2006). Effect of natural
 839 production in a zero exchange suspended microbial floc based super-intensive culture
 840 system for white shrimp *Litopenaeus vannamei*. *Aquaculture*, 258, 396-403.

841 Watten, B.J., & Busch, R.L.(1984). Tropical production of tilapia (*Sarotherodon*
842 *aurea*) and tomatoes (*Lycopersicon esculentum*) in a small-scale recirculating water
843 system. *Aquaculture*, 4, 271-283

844 White, K., O'Neill, B., & Tzankova, Z. (2004). *At a Crossroads: Will Aquaculture*
845 *Fulfil the Promise of the Blue Revolution?* SeaWeb. <http://www.seaweb.org/home.php>
846

847 Worm, B., Barbier, E. B., Beaumont, N., Duffy, E., Folke, C., Helpern, B. S.,
848 Jackson, J. B. C., Lotze, H. K., Micheli, F., Palumbi, S. R., Sala, E., Selkoe, K. A.,
849 Stachowicz, J. J., & Watson, R. (2006)., Impacts of Biodiversity Loss on Ocean
850 Ecosystem Services, *Science*, 314, 787-790.
851

852 Young, J.A., Grady, K., Little, D., Watterson, A. & Murray, F. (2006).
853 Multidisciplinary Perspectives on an Emergent Fish Product: The Tank of British
854 Tilapia. (10 pp) In: *Proceedings of the Thirteenth Biennial Conference of the*
855 *International Institute of Fisheries Economics & Trade*, July 11-14, 2006,
856 Portsmouth, UK.
857
858

859

860

861 .

862