

'Paternity analysis of pollen-mediated gene flow for *Fraxinus excelsior* L. in a chronically fragmented landscape'

By C F E Bacles and R A Ennos.

In *Heredity*, Volume 101, Issue 4, pp. 368 – 380, 2008.

DOI: 10.1038/hdy.2008.66

<http://www.nature.com/hdy/journal/v101/n4/abs/hdy200866a.html>

1 **ORIGINAL ARTICLE**

2 **Paternity analysis of pollen-mediated gene flow**
3 **for *Fraxinus excelsior* L. in a chronically**
4 **fragmented landscape**

5

6 **CFE Bacles^{1,2} and RA Ennos^{1,3}**

7

8 *¹Institute of Evolutionary Biology, Ashworth Laboratories, School of Biological*
9 *Sciences, The University of Edinburgh, Edinburgh, EH9 3JT, UK*

10 *²Present address: INRA UMR BIOGECO 69 Route d'Arcachon 33612 Cestas*
11 *cedex France*

12 *³Correspondence: Dr. Richard A. Ennos. Institute of Evolutionary Biology,*
13 *Ashworth Laboratories, School of Biological Sciences, The University of Edinburgh,*
14 *Edinburgh, EH9 3JT, UK. Email: rennos@ed.ac.uk; Phone :+44 131 650 5411 ;*
15 *Fax : +44 131 662 0478*

16

17 **Keywords:** *Fraxinus excelsior* L., habitat fragmentation, landscape genetics,
18 genetic connectivity, microsatellites, paternity assignment

19

20 **Running title:** Genetic connectivity of ash forest remnants

21

22 **Word count** (excluding references, tables and figures): 7610 words

23

24

25 Paternity analysis based on microsatellite marker genotyping was used to infer
26 contemporary genetic connectivity by pollen of three population remnants of the
27 wind-pollinated, wind-dispersed tree *Fraxinus excelsior*, in a deforested Scottish
28 landscape. By deterministically accounting for genotyping error and comparing a
29 range of assignment methods, individual-based paternity assignments were used
30 to derive population-level estimates of gene flow. Pollen immigration into a 300ha
31 landscape represents between 43% and 68% of effective pollination, mostly
32 depending on assignment method. Individual male reproductive success is
33 unequal, with 31 of 48 trees fertilising one seed or more, but only three trees
34 fertilising more than ten seeds. Spatial analysis suggests a fat-tailed pollen
35 dispersal curve with 85% of detected pollination occurring within 100m, and 15%
36 spreading between 300m and 1900m from the source. Identification of immigrating
37 pollen sourced from two neighbouring remnants indicates further effective dispersal
38 at 2900m. Pollen exchange among remnants is driven by population size rather
39 than geographic distance, with larger remnants acting predominantly as pollen
40 donors, and smaller remnants as pollen recipients. Enhanced wind dispersal of
41 pollen in a barren landscape ensures that the seed produced within the catchment
42 includes genetic material from a wide geographic area. However, gene flow
43 estimates based on analysis of non-dispersed seeds were shown to underestimate
44 realised gene immigration into the remnants by a factor of two suggesting that
45 predictive landscape conservation requires integrated estimates of post-recruitment
46 gene flow occurring via both pollen and seed.

47 **Introduction**

48 The accuracy of our prediction of the response of forest trees to deforestation and
49 population fragmentation relies on an understanding of how pollen and seed
50 movement is modified as a consequence of changes in the landscape (Sork and
51 Smouse, 2006). Trees, which are characterised by their individual longevity, high
52 intra-population genetic diversity, and often substantial potential for gene flow via
53 pollen and seed, may be particularly well-equipped to withstand habitat disturbance
54 (Hamrick, 2004). Although theoretical predictions of reduced genetic diversity and
55 elevated inbreeding following habitat fragmentation (Young *et al.*, 1996) are upheld
56 for a number of wind-pollinated temperate tree species (Sork *et al.*, 2002; Jump
57 and Penuelas, 2006), a recent review of empirical studies conducted in neotropical
58 tree species suggests that fragmentation generally has more complex effects
59 (Lowe *et al.*, 2005).

60
61 An emerging picture is of an increase in pollen and seed-mediated gene flow
62 across deforested landscapes (Aldrich and Hamrick, 1998; Dick, 2001; White *et al.*,
63 2002; Bittencourt and Stebbenn, 2007) However this enhanced gene flow does not
64 necessarily lead to an increase in genetic diversity or reduction in inbreeding if a
65 limited number of pollen and seed sources contribute to the gene pool (Aldrich and
66 Hamrick, 1998; O'Connell *et al.*, 2006; Sork *et al.*, 2002). Moreover the effect of
67 fragmentation is often not uniform over the landscape. Smaller fragments tend to
68 receive proportionally more pollen immigration than larger fragments because of a
69 paucity of local pollen donors (Sork and Smouse, 2006). It is clear from these
70 considerations that to understand the genetic connectivity of tree species living in
71 fragmented habitats requires an appreciation of the contemporary processes of
72 dispersal and establishment and an analysis of how they are affected by the spatial

73 scale of fragmentation and the heterogeneity of the landscape in which the
74 fragmentation occurs (Sork and Smouse, 2006).

75
76 The combined development of highly polymorphic microsatellite markers and
77 statistical analysis of parentage assignment (reviewed in Jones and Ardren, 2003)
78 have made it possible to gather empirical evidence of contemporary gene
79 movement within various landscapes for wild animals (e.g. Hazlitt *et al.*, 2006) and
80 plants (e.g. Bittencourt and Stebbenn, 2007). For measuring contemporary gene
81 flow these methods have many advantages over previous approaches that relied
82 on inferences from genetic structure and yielded estimates of historical, average
83 values of effective migration rate (Sork *et al.*, 1999; Whitlock and McCauley, 1999).
84 However a number of problems are beginning to emerge with adopting parentage
85 assignment approaches for measuring gene flow in practice.

86

87 The first issue is that the microsatellite methodology widely used for genotyping
88 has significant assay limitations that may call the accuracy of the pedigree
89 inference into question (Dakin and Avise, 2004; Hoffman and Amos, 2005; Slavov
90 *et al.*, 2005). Two main types of genotyping error can be distinguished, allele
91 dropouts (Dakin and Avise, 2004) and erroneous calling of allele size (Amos *et al.*,
92 2007) both of which can be either of a systematic or stochastic nature. While some
93 studies suggest that it is best to discard affected loci from parentage analyses
94 where occurrence of non-amplifying (null) alleles is suspected (Dakin and Avise,
95 2004), Wagner *et al.* (2006) argue that when the number of loci is low,
96 discriminatory power may decrease dramatically as a result. They suggest that a
97 better alternative to either removing loci or ignoring the presence of null alleles is to
98 accommodate them within the analyses. Indeed the use of many, even moderately
99 variable, loci rather than fewer hypervariable ones, reduces the impact of error at

100 any particular locus on parentage assignment (Hoffman and Amos, 2005; Slavov *et*
101 *al.*, 2005). Despite the profusion of recent publications establishing that even a low
102 genotyping error rate had non-trivial consequences for parentage and relatedness
103 studies, quantification and publication of error rates are not yet routinely performed
104 (Hoffman and Amos, 2005).

105

106 The second issue is that conclusions drawn from the analysis depend on the
107 method of paternity assignment adopted and assumptions about the size and
108 genetic composition of the population of potential paternal parents that have not
109 been sampled (Oddou-Muratorio *et al.*, 2003; Burczyk and Chybicki, 2004). While
110 simple exclusion is a useful starting point for paternity inference, refined statistical
111 approaches are necessary to assess the confidence in paternity assignment
112 (Marshall *et al.*, 1998). For example in natural tree populations it is virtually
113 impossible to sample all trees contributing to the reproductive pollen pool. It is
114 therefore necessary to assess the risk of excluding a candidate pollen parent on
115 the sole grounds that it has not been sampled. Recent methods based on either
116 Likelihood or Bayesian approximation allow us to estimate the statistical precision
117 of a paternity assignment for a given sample of a reproductive population (Marshall
118 *et al.*, 1998; Nielsen *et al.*, 2001; Gerber *et al.*, 2003; Araki and Blouin, 2005;
119 Hadfield *et al.*, 2006). Overall, it is preferable to use more than one of these
120 approaches to estimate genetic exchange among populations (Oddou-Muratorio *et*
121 *al.*, 2003).

122

123 In plants a further complication with measuring interpopulation gene flow using
124 parentage assignment arises because gene flow is effected by two asynchronous
125 dispersal processes, the first involving pollen and the second involving seeds. The
126 genetic material transferred between populations via pollen and incorporated into

127 seed present on a mother tree only brings about gene flow if that seed is recruited
128 into the local population. Parentage assignment of recruited seedlings is difficult to
129 track with current analytical tools (Sork and Smouse, 2006). Therefore
130 contemporary gene flow among fragmented tree populations has often been
131 estimated by measuring pollen movement to pre-recruitment seeds. These
132 estimates of gene flow are reasonable if seed-mediated gene flow among existing
133 populations is rare, and seed dispersal is primarily important for recolonisation and
134 range expansion. However in tree species with a high potential for long-distance
135 seed dispersal and subsequent recruitment this assumption may be invalid
136 (Smouse and Sork, 2004) and realised gene flow following seed dispersal may
137 differ significantly from gene flow measured in pre-recruitment seed within a
138 population. The extent of the discrepancy between gene flow measured from
139 parentage analysis before and after seed recruitment has still to be properly
140 documented.

141

142 The overall objective of the current study is to describe the genetic connectivity
143 among population fragments of common ash (*Fraxinus excelsior*) in a chronically
144 deforested landscape in the Southern Uplands of Scotland. Previous work on these
145 fragments has shown that they maintain high levels of genetic diversity and weak
146 inter-fragment differentiation ($\theta = 0.080$), indicating that historical gene flow has
147 not been limited ($Nm = 3.48$). We also found from an analysis of seed families,
148 using a mixed-mating model approach (Ritland 2002), that contemporary matings
149 are, on average, predominantly outcrossed ($t_m = 0.971 \pm 0.028$) and using a
150 neighbourhood model approach (Burczyk *et al.*, 2002) that contemporary effective
151 pollen dispersal distance within the landscape averages 328 m (Bacles *et al.*,
152 2005). Both seeds collected from forest fragments and newly recruited seedlings
153 were found to harbour high levels of genetic diversity comparable to that of the

154 adult population suggesting an essential contribution of long distance dispersal to
155 genetic diversity in this wind-pollinated wind-dispersed species (Bacles *et al.*, 2005;
156 2006). The present paper complements these studies by quantifying the genetic
157 exchange among the individual fragments brought about by pollen flow and relates
158 this to the size and landscape context of the fragments.

159

160 To achieve this we estimate pollen-mediated gene flow and male reproductive
161 success of local *F. excelsior* trees from a paternity analysis of non-dispersed seeds
162 genotyped at hypervariable microsatellite markers. This is the best methodological
163 approach currently available to address this question. Nonetheless, in full
164 awareness of the potential limitations of the methodology, we take a number of
165 steps in order to ensure that our estimation describes true biological phenomena.
166 Firstly, we quantify genotyping error at marker loci and set out to minimise error
167 due to possible mis-scoring or null alleles using a simple deterministic approach.
168 Secondly, we use a range of paternity assignment methods to obtain a confidence
169 interval rather than a point estimate of pollen-mediated gene flow. Lastly, we
170 compare such estimates derived from non-dispersed seeds with those derived from
171 seedlings establishing in the same *F. excelsior* remnants estimated by means of
172 parent-pair analysis (Bacles *et al.*, 2006), to assess how they relate to absolute
173 levels of genetic exchange among remnant populations.

174 **Material and methods**

175 **Study species**

176 *Fraxinus excelsior*, common ash, is a post-pioneer tree species widespread in
177 temperate Europe and native throughout the British Isles. The phylogeography of
178 the species is now well described (Ferrazzini *et al.*, 2007; Heuertz *et al.*, 2004;
179 Morand *et al.*, 2002). *F. excelsior* displays a complex, polygamous sexual system
180 (FRAXIGEN, 2005) in which individuals may be classified phenotypically across a
181 continuum from purely male to purely female with a whole range of hermaphroditic
182 intermediates. Hermaphrodite individuals are self-fertile and levels of seed sets are
183 similar in hermaphrodite and female trees, but in natural populations, *F. excelsior* is
184 preferentially outcrossed and male fertility of hermaphrodite trees appears to be
185 much lower than that of male trees (FRAXIGEN, 2005). Fruits are dry and winged,
186 adapted to wind-dispersal. Regular fruit bearing begins around 20 years of age but
187 fruiting phenology will vary depending on latitude, altitude, temperature and
188 between years with great variation from no seeding to masting (FRAXIGEN, 2005).

189

190 **Sampling and data collection**

191 The study site is a highly deforested catchment of 900ha (Moffat Dale) located
192 80km south of Edinburgh, Scotland (N 55° 23' 51" W 3° 19' 50"), which forms part
193 of a glacially derived landscape in which steep sided valleys have been carved by
194 ice. Many native tree species including *F. excelsior* are confined to steep and
195 narrow stream sides situated at the bottom of steep valleys inaccessible to grazing
196 herbivores. Populations of *F. excelsior* tend to be very small, comprising ten to 30
197 mature individuals, with no natural regeneration in grazed areas. Remnant stands
198 are typically separated by hundreds of metres although some can be isolated by
199 more than one kilometre. In this catchment, *F. excelsior* is present in only five

Figure 1

200 forest remnants, two of them within the Carrifran Burn and three others in its
201 immediate surroundings (Figure 1, see also Bacles *et al.* 2005).

202

203 Two remnants in the bare open landscape of the Carrifran Burn, CDa and CMa,
204 and one remnant confined to a higher altitude dense conifer plantation upstream of
205 Carrifran in Swine Cleuchs, SCa, chosen for their heterogeneity in size, density
206 and landscape features, were exhaustively sampled for adult trees and family
207 arrays. Two neighbouring remnants, in Spoon Burn (SBa), the adjacent valley to
208 Moffat Dale nearest to Carrifran, downstream and in Whitewells (Wa) located at the
209 bottom of the Moffat Dale where the Carrifran streams drain into Moffat Water
210 (Figure 1), were partially sampled for adult trees to gather an indication of potential
211 pollen immigration because they are the only two other known local pollen sources
212 within ten km.

213

214 In 2000 and 2001, leaf material was collected from all mature trees in CDa, CMa
215 and SCa (comprising 30, four and 12 individuals respectively, Figure 1). Leaf
216 material was also collected from two trees (A, B) isolated from the nearest
217 remnants by a distance of 250 m and from a sample of 20 mature trees in SBa and
218 Wa (Figure 1). Such sampling represents no less than 40% of the composition of
219 these two remnants. 30 fruits, or all fruits if the seed crop was less, were collected
220 from all 19 trees producing fruits in 2000, a non-masting year, in each of remnants
221 CDa, CMa and SCa. In total, we sampled 88 trees and 483 seeds from 19 families.

222

223 The complete sample of 88 trees and 483 seeds were genotyped for five
224 microsatellite markers previously developed for *F. excelsior*, namely, M2-30B, 1.19
225 and 3.1 (Brachet *et al.*, 1999), and FEMSATL2 and FEMSATL5 (Lefort *et al.*, 1999).
226 DNA isolation, amplification by polymerase chain reaction (PCR) and

227 electrophoretic separation of PCR products were carried out as described
228 elsewhere (Bacles *et al.*, 2005).

229

230 **Evaluating microsatellite scoring and accounting for mistyping**

231 Out of 483 seeds genotyped, 61 presented a mismatch with their mother at one or
232 more loci. Furthermore, genotypes observed at loci 3.1 and FEMSATL 5 suggest
233 departure from Mendelian segregation (Bacles *et al.*, 2005) and occurrence of null
234 alleles which has also been discussed in other *F. excelsior* studies (e.g. Heuertz *et*
235 *al.*, 2001; Morand *et al.*, 2002).

236

237 In the rare instances where correction for genotyping error is applied in empirical
238 studies, it is generally introduced as a global stochastic error rate (Gerber *et al.*,
239 2000; Marshall *et al.*, 1998). A major drawback of such practice is that the benefits
240 of accounting for error are often outweighed by costs in precision of paternity
241 assignment which becomes uninformative (Oddou-Muratorio *et al.*, 2003; Morissey
242 and Wilson, 2005). Therefore, here we chose to deterministically account for allele
243 dropouts and size miscalls by performing two successive transformations to the
244 raw multilocus genotypes (referred to hereafter as RAW).

245

246 Erroneous allele sizing is most likely to occur between alleles of similar size and
247 when alleles are rare. Therefore, we applied an initial transformation to account for
248 size miscalls in the form of a binning procedure. At each locus, rare alleles were
249 binned with common alleles of the nearest size. Alleles were deemed rare when
250 they occurred at a frequency of less than 0.01 for the entire dataset. The procedure
251 was applied strictly to loci M2.30B and FEMSATL2. For loci 1.19, 3.1 and FEMSATL5,
252 more common alleles were also binned in order to reflect difficulties in gel scoring
253 for 1.19 and difficulties in respect of Mendelian segregation for 3.1 and FEMSATL5

254 (Bacles *et al.* 2005). The procedure reduced the number of alleles observed from
255 54 to 29, from 35 to 20, from 10 to 7, from 46 to 12 and from 17 to 8 at loci
256 FEMSATL2, M2-30B, 1.19, FEMSATL5 and 3.1 respectively in the binned dataset
257 (referred to hereafter as BIN).

258

259 A second transformation was then performed to account for the possible
260 occurrence of allele dropouts. A one-allele dropout model was applied to each
261 locus by introducing a new (i.e. non-observed) allele, by rescored every individual
262 with a non-amplifying genotype as homozygote null, and every observed
263 homozygote as heterozygote null in the transformed dataset (hereafter referred to
264 as BINNULL).

265

266 For each dataset, genotyping error rates were quantified by means of direct
267 comparison of offspring-mother genotype at each locus and averaged over loci in
268 CERVUS 2.0 (Marshall *et al.*, 1998). Loci were retained for subsequent analyses if
269 the estimated genotyping error was less than 5%. In order to assess the
270 discriminatory power of each dataset, a paternity exclusion probability (PEP) was
271 computed for each locus and accumulated over loci in FAMOZ (version released on
272 17.04.2007, Gerber *et al.*, 2003).

273

274 **Estimating contemporary pollen-mediated gene flow at the landscape scale**

275 Paternity analyses were undertaken to identify the pollen parent of the 422 seeds
276 that shared a compatible multilocus genotype with their putative mother only and
277 excluding the 61 seeds presenting at least one mismatching allele with their
278 mother. Pollen parents were considered either among the 48 trees sampled in
279 CDa, CMa and SCa, including the possibility for mother trees to self, or as pollen-

280 mediated gene flow from outside the landscape covered by the three completely
281 censused remnants (Figure 1).

282

283 For each of the RAW, BIN and BINNULL datasets, paternity was assigned using both
284 a simple exclusion (SE) and a maximum-likelihood (ML) approach in FAMOZ (Gerber
285 *et al.*, 2003). In each case, outcomes of paternity assignment may be, for each
286 individual seed, either that one unique individual among the 48 trees of Carrifran
287 and Swine Cleuchs is assigned as its pollen parent, or that its paternity is
288 unresolved with more than one possible pollen parent among the 48 trees, or
289 finally that all 48 trees are excluded as potential pollen parents and its paternity is
290 assigned to immigrant pollen. A range of values for apparent pollen-mediated gene
291 flow into the landscape is subsequently obtained as the percentage of seeds in the
292 sample for which paternity was assigned to immigrant pollen.

293

294 In FAMOZ, confidence in paternity assignment is estimated using a simulation
295 procedure for hypothesis testing (Gerber *et al.*, 2000). The paternity of the 422
296 seeds sampled from 19 mother trees in three *F. excelsior* remnants of Moffat Dale
297 was assigned to the most-likely fathers detected by means of 'log of the odds'
298 ratios (LOD scores) based on pollen pool gene frequencies estimated from progeny
299 arrays in MLTR (Ritland, 2002). We chose to approximate the (non-observed) allele
300 frequencies of the reproductive population by using the observed pollen pool
301 frequencies instead of the frequencies observed for the small sample of 88 mature
302 trees because the latter, which is sampled *a priori* based on spatial location, may
303 be a poor estimate of the actual reproductive population if gene flow is extensive.
304 No significant genotypic association was detected among any pair of loci (Bacles *et*
305 *al.*, 2005). LOD scores over all loci were therefore obtained by adding LOD scores
306 calculated for each locus.

307

308 Confidence in paternity assignment was then determined in FAMOZ by comparing
309 the distribution of the LOD scores of the most-likely fathers of 50 000 randomly
310 generated seeds with their father randomly chosen among the 48 trees to the
311 distribution of LOD scores of the most-likely fathers of 50 000 seeds whose paternal
312 genotype was randomly generated according to pollen pool allele frequencies. The
313 test threshold for rejecting a candidate as a true father (TF), was chosen at the
314 intersection of the two distributions of LOD scores to minimise both type I error,
315 wrongly considering as resulting from pollen immigration a seed sired by a sampled
316 father, and type II error, wrongly assigning true pollen immigration to a sampled
317 father (Gerber *et al.*, 2000). For paternity assignment by simple exclusion (SE), all
318 candidate males with a positive LOD score (i.e. test threshold TRUE FATHER=0) were
319 not excluded from being the true father.

320

321 Global results of paternity assignment obtained for each of RAW, BIN and BINNULL
322 datasets with both SE and ML methods are discussed in respect of estimated error
323 rates, confidence levels in assignments and estimated pollen-mediated gene flow
324 at the landscape scale. The dataset/method combination found to minimise
325 genotyping error rates while maximising confidence in paternity assignments was
326 retained for subsequent detailed description of individual male reproductive
327 success. In particular, results were summarised in order to identify the number of
328 sires among the sample trees and the number of seeds they sired among the
329 sampled seeds. The pollen dispersal curve was estimated by plotting the distance
330 between mother trees and pollen donors for each most-likely assignment. When
331 more than one likely father was identified (unresolved assignment), a fraction of the
332 seed was assigned to all likely fathers evenly and proportionally to the number of
333 likely fathers found.

334 **Estimating the fractional pollen contribution of forest remnants and**
335 **identifying local sources of pollen immigration**

336 In order to estimate landscape connectivity and pollen-mediated genetic exchange
337 among forest remnants, it may be most relevant to assess the relative pollen
338 contribution of forest remnants to the seed crop rather than to define individual
339 paternity *per se*. It has been argued that such (meta)population-scale phenomena
340 may be better addressed with fractional-likelihood assignment methodology that
341 will assign a fraction of the paternity of a given seed to all male candidates with a
342 positive LOD score in proportion to their likelihood probability (Nielsen *et al.*, 2001).
343 In order to estimate potential pollen immigration into CDa, CMa, and SCa from
344 other known *F. excelsior* remnants of Moffat Dale, SBa and Wa (Figure 1), we
345 estimated the posterior expectation of the number of sampled offspring in each of
346 five remnants by means of fractional-likelihood assignment in PATRI (Signorovitch
347 and Nielsen, 2002).

348
349 The approach in PATRI also allows us to make prior assumptions about the
350 proportion of the pollen parents that have not been sampled (Nielsen *et al.*, 2001).
351 While the actual effective male population size is unknown, we do have some
352 expectations of the number of trees occurring in the landscape and likely to
353 contribute to the pollen pool. Therefore, we tested the sensitivity of the fractional-
354 likelihood assignment to assumptions made on the population size (N) by
355 successively repeating analyses for an N of 88, the number of trees sampled; an N
356 of 150, the approximate number of *F. excelsior* trees occurring in the catchment
357 and N modeled as a uniform function varying between 100 and 500. We compared
358 results with those of a maximum-likelihood assignment performed in FAMOZ when
359 considering all 88 trees sampled in Moffat Dale.
360

361 **Comparing potential to realised pollen-mediated gene flow among forest**
362 **remnants.**

363 To assess whether estimates of pollen-mediated gene flow from seeds that have
364 not yet dispersed reflect estimates of pollen-mediated gene flow seen after seed
365 dispersal and establishment, for each of remnant CDa, CMa and SCa, we used
366 most-likely fathers to attribute the origin of the pollen grain to either local pollen,
367 foreign pollen of known source (in other identified remnants) or of unknown source.
368 We compared these figures with previously published results derived from a ML
369 parent-pair analysis performed on seedlings establishing in the same three
370 remnants (Bacles *et al.*, 2006).

371

372 In addition we estimated total gene flow into fragments using genotypic data
373 generated both from progeny arrays (T_p) and from established seedlings (T'_s). Note
374 that pollen grains only carry one gene copy while diploid seeds carry two. Let A
375 and A' represents the number of local seeds fertilised by immigrant pollen in
376 progeny arrays and establishing seedlings respectively. Let B' be the number of
377 immigrating seeds, and C and C' the total number of seeds sampled in progeny
378 arrays and establishing seedlings respectively, then:

379

380 $T_p = (A / 2C) \times 100$ **Equation 1**

381 and

382 $T'_s = (A' + 2B') / 2C' \times 100$ **Equation 2**

383

384 In equation 1, seeds are sampled on known mother trees and are all local. T_p
385 therefore assumes that pollen is the main vector of gene flow among populations
386 and that seed dispersal is mostly local. Results are discussed in terms of

Published in *Heredity* 101 (2008) : 368-380.

387 comparison of potential (i.e. *ante* dispersal) and realised (i.e. *post* dispersal and
388 establishment) gene flow in the three heterogeneous forest remnants.

389 **Results**

390 **Genotyping error and choice of dataset**

391 Genotyping error rates estimated by means of offspring-mother genotype
392 comparison are reported for each locus and each genotype transformation in Table
393 1. They were found to be highest for loci 3.1 (error (RAW) = 0.2847) and FEMSATL 5
394 (error (RAW) = 0.2237). Drastically reducing the number of alleles, from 17 to 8 and
395 from 46 to 12 at locus 3.1 and FEMSATL5 respectively, decreases the error rate only
396 slightly (error (3.1, BIN) = 0.2679; error (FEMSATL5, BIN) = 0.2133) but additional
397 inclusion of a null allele decreases the error significantly (error (3.1, BINNULL) =
398 0.1499; error (FEMSATL5, BINNULL) = 0.0762) suggesting that null alleles may be
399 responsible for the non-Mendelian segregation observed in progeny arrays at
400 these loci. In contrast, genotyping error rates estimated at loci 1.19, M2.30B and
401 FEMSATL2 were under 5% (Table 1), and lowest at loci M2.30B and FEMSATL2 after
402 binning of alleles (error (M2.30B, BIN) = 0.0347; error (FEMSATL2, BIN) = 0.0443) and
403 with inclusion of a null allele at locus 1.19 (error (1.19, BINNULL) = 0.0289). Overall,
404 inclusion of loci 3.1 and FEMSATL5 in estimates increases mean error across loci
405 dramatically, up to 3 times (Table 1). When these loci were excluded, estimates of
406 mean error across loci were consistently under 5%.

407
408 Paternity exclusion probabilities (PEP) estimated per locus for RAW vary between
409 0.594 at locus 1.19 and 0.864 at locus FEMSATL2 (Table 1) reflecting variation in
410 level and evenness of polymorphism among loci (Bacles *et al.*, 2005). As expected,
411 reducing the number of alleles at each locus systematically lowers discriminatory
412 power among genotypes at each locus, albeit moderately, with PEP estimated per
413 locus from BIN varying from 0.504 at locus 1.19 to 0.857 at locus FEMSATL2.
414 Conversely, additional inclusion of a null allele results in contrasting effects on
415 single locus estimates of PEP (Table 1). However, cumulative estimates of PEP,

Table 1

416 including all five loci, are consistently very high across datasets, reaching values
417 upward of 99.9%. Excluding loci 3.1 and FEMSATL5 decreased cumulative PEP only
418 slightly (Table 1).

419

420 On the basis both that including loci 3.1 and FEMSATL5 increases genotyping error
421 to rates well above 5% and that excluding them hardly affects multilocus genotype
422 discrimination of individuals, they were not retained for subsequent analyses.

423 Meanwhile, considering loci 1.19, M2.30B and FEMSATL2 only, mean genotyping
424 error rate is lowest when multilocus genotypes are transformed with binning at loci
425 M2.30B and FEMSATL2 and with binning and inclusion of a null allele at locus 1.19
426 (dataset hereafter referred to as BIN3NULL1; error (MEAN, BIN3NULL1) = 0.0360). No
427 identical multilocus genotype was found among the 88 trees sampled in Moffat
428 Dale and cumulative PEP is estimated at 0.991 which is nearly as high as for RAW
429 data (Table 1).

430

431 **Contemporary pollen-mediated gene flow at the landscape scale**

432 Results of SE and ML paternity analysis for RAW, BIN, BIN3NULL1 datasets and for
433 assignment of pollen parents to the 422 seeds that do not show any mismatch with
434 their mother on raw data, are given in table 2. 43% to 68% of the 422 seeds
435 analysed were found to have been fertilized with pollen dispersed from trees
436 located outside the landscape covered by three *F. excelsior* remnants of Carrifran
437 and Swine Cleuchs. Highest estimates were obtained using a ML method (Table 2).
438 Differences between SE and ML estimates of apparent pollen flow are due to a
439 number of seeds that were not assigned a father among the 48 sampled trees with
440 the ML method because their LOD was positive but below the given threshold for
441 assignment. However they were assigned one (or several potential) fathers with
442 the SE method, most frequently for the bin dataset ($N_{\text{assigned}}(\text{BIN, SE}) = 239$, Table

Table 2

443 2). False rejection of true sampled fathers was lowest (Type I error < 0.05 for TF =
444 2.90) for fewer seeds ($N_{\text{assigned}} = 141$) with the BIN3NULL1 transformation (Table 2).

445

446 **Individual male reproductive success and distance of pollen dispersal**

447 On the basis that the BIN3NULL1 dataset is characterised by the lowest estimates of
448 genotyping error and type I error in ML assignment, the most-likely pollen parents
449 identified for 141 seeds under these conditions were retained for subsequent
450 description of male reproductive success and spatial patterns of pollen dispersal at
451 the landscape scale.

452

453 In total, 31 of the 48 trees sampled in CDa, CMa and SCa were found to sire one
454 or more of the sampled seeds (Supplementary Table S1). Of these, 14 trees
455 fertilised three seeds or less, while three trees fertilised more than ten seeds each.
456 In the latter case, all fertilized seeds were sampled from neighbouring trees within
457 the same remnant as the sire.

458

459 This pattern in individual male reproductive success is reflected in the L-shape of
460 the pollen dispersal curve constructed by plotting the distribution of spatial
461 distances between the 141 assigned seeds (i.e. based on spatial location of
462 mother trees) and their most-likely pollen-parent (Figure 2a). Over 80% of effective
463 pollen dispersal is confined to less than 100m from the source, corresponding to
464 local pollen movement within remnant. The observed proportion of these local
465 pollinations is significantly higher than expected under random dispersal (48%,
466 Wilcoxon two-sided signed rank test p -value<0.05, Figure 2a). However, a number
467 of rarer events, each representing less than 10% of total effective pollen dispersal
468 (which is up to 25% less than expected from random dispersal), were identified

Figure 2

469 between 200m and 600 m, and between 1600m and 1800m corresponding to inter-
470 remnant long-distance dispersal.

471

472 Further analysis that included trees in partially sampled remnants SBa and Wa
473 allowed us to refine the shape of the pollen dispersal distribution (Figure 2b). ML
474 paternity assignment of 163 seeds ($TF=2.92$, Type I error <0.05 ; Type II error <0.28)
475 attributed their paternity either to trees sampled in CDa, CMa and SCa or to trees
476 sampled in SBa and Wa (Supplementary Table S2). The inclusion of these nearby
477 pollen sources allowed the detection of a small number of effective pollination
478 events at distances between 1100m and 1600m, between 1800m and 200m and
479 as great as 2900m (Figure 2b).

480

481 **Pollen-mediated genetic connectivity of forest remnants**

482 Pollen contribution of the five *F. excelsior* population remnants sampled for adult
483 trees was estimated from fractional likelihood (FL) paternity analysis of a subset of
484 404 seeds and 83 male candidates with no missing value in their multilocus
485 genotype at 1.19, M2.30B and FEMSATL2 because PATRI computes complete
486 genotypes only (Signorovitch and Nielsen, 2002). Estimates of the relative
487 contribution of the five *F. excelsior* remnants of Moffat Dale to effective pollination

Table 3 488 in remnants CDa, CMa and SCa are similar when a FL or ML approach is applied to
489 all 88 genotyped trees (Table 3). However, the absolute fractional contribution of
490 remnants to paternity of sampled seeds is highly sensitive to the sampled fraction
491 of reproductive adults (Nielsen *et al.* 2001) as illustrated by decreasing posterior
492 expectation of the number of sampled offspring fathered in each remnant with
493 increasing prior N (Table 3). Overall however, the relative pollen contribution of the
494 five forest remnants remains unchanged. CDa contributes most to paternity of the
495 sampled seeds. SCa, SBa and Wa also contribute in decreasing proportion (Table

496 3), while remnant CMa is a poor contributor (Table 3). The analysis clearly
497 demonstrates that neighbouring remnants may act as sources of immigrant pollen.

498

499 Estimates of pollen-mediated genetic exchange among remnants CDa, CMa and
500 SCa derived from ML paternity analysis confirm that the largest remnant CDa acts
501 as a pollen donor, siring 26% of the seeds sampled in remnant CMa, located 600m
502 away, and 7% of the seeds sampled in the most spatially isolated remnant, SCa,
503 located 1700m away (Figure 3). Conversely, CMa, the smallest remnant ($N_{trees} =$
504 4), only sired 3% of its local seeds, and 1% the seeds within CDa (Figure 3).

Figure 3

505

506 **Potential and realised pollen-mediated gene flow**

507 How such pollen-mediated genetic exchange will impact on genetic structure
508 depends on dispersal and establishment of the seeds. Estimates of potential
509 pollen-mediated gene flow into remnants CDa, CMa, and SCa from ML paternity
510 analysis of 422 seeds collected on mother trees before their dispersal described
511 above (65% to 94%, Table 4) are comparable to those of potential pollen-mediated
512 gene flow from a ML parent-pair analysis of 60 seedlings that were establishing in
513 the same three remnants the following year (70% to 100%, Table 4). However,
514 such comparison also show that pollen-mediated gene flow realised after seed
515 dispersal and seedling establishment is much lower, ranging from 12.5% in
516 remnant CMa to 17.5% in remnants CDa and SCa. Total gene flow estimates from
517 progeny arrays are much lower (T_p ranging between 32.5% and 47%) than from
518 establishing seedlings (T 's ranging between 67.5% and 87.5%, Table 4).

Table 4

519 **Discussion**

520 Despite a number of significant concerns over genotyping error and uncertainties
521 associated with statistical modeling, the application of paternity assignment
522 analysis in the fragmented populations of *F. excelsior* has significantly enhanced
523 our understanding of their genetic behaviour. It is clear that the population
524 fragments within a single valley receive about half their pollen from outside the
525 valley. Remnants within the valley are genetically connected via pollen flow, but the
526 patterns of pollen flow among fragments are not symmetrical; pollen is
527 preferentially transferred from large to small fragments. The analysis has also
528 demonstrated that the effective pollen dispersal curve is fat-tailed. While the
529 majority of detected pollen movement occurs over short distances (within 100m),
530 there is still substantial pollen flow occurring over distances greater than 1km.
531 Although these general conclusions are important for guiding the management of
532 fragmented tree populations, this study has also highlighted the practical difficulties
533 associated with obtaining quantitative assessments of gene flow from large scale
534 studies that rely on parentage analysis.

535

536 A predictive understanding of the genetic connectivity of fragmented populations
537 requires reliable estimation of contemporary gene dispersal across heterogeneous
538 landscapes (Sork and Smouse, 2006). While development of both molecular
539 techniques and statistical tools has greatly improved prospects for accuracy, the
540 application of parentage analyses to natural populations remains an evolving area
541 of research leading to regular reanalysis of empirical data within new statistical
542 frameworks (e.g. Hadfield *et al.*, 2006; Slate *et al.*, 2000). At the centre of the
543 debate, lies the question of sensitivity of parentage analyses to partial sampling of
544 the reproductive population and to genotyping error at marker loci (Nielsen *et al.*,
545 2001; Oddou-Muratorio *et al.*, 2003; Slavov *et al.*, 2005). In order to obtain reliable

546 population-level inference of gene flow from a collection of individual-level paternity
547 assignments, we chose to address these concerns by applying a range of paternity
548 analysis methods to *F. excelsior* population remnants of the chronically deforested
549 catchment of Moffat Dale (Table 2 and 3).

550

551 Critically, application of parentage analyses to estimating gene movement in
552 natural populations relies on a conundrum: accuracy in estimation of the proportion
553 of the reproductive population that has not been sampled (i.e. immigrant gene flow)
554 strongly increases as the proportion of the -yet unknown- reproductive population
555 that has not been sampled decreases (Oddou-Muratorio *et al.*, 2003). Approaching
556 true reproductive population size seems particularly important to analyses
557 performed when using a fractional likelihood approach in PATRI because estimating
558 the absolute contribution of *F. excelsior* trees to paternity of sampled seeds is
559 sensitive to input prior information on the effective male population size (Table 3).

560 The advantage of the hypothesis testing-based simulation approach to determining
561 assignment confidence in FAMOZ is that it does not require any assumption on the
562 size of the true reproductive population. Comparison of SE and ML methods for a
563 range of transformed multilocus genotypes accounting for genotyping error *sensu*
564 *lato* at microsatellite markers suggest an immigration of at least 43% (SE, BIN) and
565 up to 68% (ML, RAW) of the pollen fertilising seeds from 19 trees of three forest
566 remnants of the Moffat Dale catchment.

567

568 The range in gene flow rates seems mostly affected by the choice of paternity
569 assignment method rather than by dataset transformation. Indeed, while
570 transformation of raw data allowed us both to reduce mean genotyping error (for
571 BIN3NULL1 at 3.60%), and to minimise false rejection of true fathers that were
572 sampled (for BIN3NULL1 Type I < 5%) to acceptable levels, estimates of gene flow

573 between RAW, BIN and BIN3NULL1 vary by up to 15% or a given paternity
574 assignment method. Variation in gene flow estimates between the SE and ML
575 methods can be attributed to the fact that, under ML, between 10% and 16% of
576 seeds are not assigned a father among the sampled trees. This is because the LOD
577 score of candidates is positive but below the given threshold for assignment
578 ($TF=2.90$, Table 2). Although paternity of these seeds cannot be assigned at the
579 chosen confidence threshold, it is arguable that their paternity should necessarily
580 be attributed to immigrant pollen. Indeed, it has been demonstrated that
581 assignment error may be much higher than random on unobserved (i.e. immigrant)
582 events (Slate *et al.*, 2000) which suggests that estimates of pollen-mediated gene
583 flow from ML method (here inclusive of seeds that were not assigned a father
584 because genotypically compatible candidates had a low LOD score) should be seen
585 as upper limits. Conversely, ML analysis suggests that even for a strict LOD score
586 threshold, Type II error of wrongly assigning immigrant pollen to an unrelated
587 sampled tree is high (up to 27% for BIN and $TF=2.50$, Table 2) indicating substantial
588 cryptic gene flow. Therefore, apparent pollen flow estimated with relaxed
589 assignment (equivalent to $TF>0$) by SE method, and those obtained with BIN data
590 because allele binning results in lower discrimination of multilocus genotypes, are
591 most conservative, with increased risk of cryptic gene flow and therefore represent
592 lower limits of effective pollen immigration into forest remnants of Moffat Dale.

593

594 We justify our deterministic transformation of genotypic data not as substitution of
595 raw datasets for transformed ones that may be more biased but rather as a simple
596 way of minimising genotyping error and its possible influence on paternity
597 assignment. The transformed dataset still includes a mean genotyping error rate of
598 about 3.6% per locus which may still have an impact on the conclusions drawn
599 from this study. Nonetheless, we deliberately chose not to include this global rate in

600 paternity analyses because there is evidence that including global genotyping error
601 rates inflates errors in paternity assignments (Oddou-Muratorio *et al.*, 2003; Slavov
602 *et al.*, 2005). Given the limitations of the dataset, which are here clearly quantified,
603 a range of estimates from several paternity analyses provides an ecologically
604 meaningful interpretation of pollen-mediated gene flow at the landscape scale.

605

606 Extensive contemporary pollen-mediated gene flow averaging 60% has already
607 been reported in plots located within continuous stands of *F. excelsior* (Hebel *et al.*,
608 2007) and of other wind-pollinated temperate tree species, for instance for *Quercus*
609 (Dow and Ashley, 1998) covering only small areas. Contemporary pollen-mediated
610 gene flow estimates of 43% to 68% for *F. excelsior* in the fragmented landscape of
611 Moffat Dale are comparatively higher because all *F. excelsior* trees were sampled
612 in an area of 300ha suggesting that *F. excelsior* maintains extensive pollen
613 exchange across a landscape heavily deforested not only locally but also at the
614 wider regional scale of the Southern Uplands of Scotland (>50km). Although trees
615 standing solitarily in grazed pastures, spatially isolated from congeners following
616 deforestation, were once described as *living-dead* (Janzen, 1986), there is now
617 plethora of evidence of reproductive activity of isolated pasture trees, mainly in
618 tropical species (Aldrich and Hamrick, 1998; Dick, 2001; White *et al.*, 2002)
619 corroborating our findings of enhanced pollen-mediated gene flow following
620 anthropogenic disturbance. Nonetheless, of the two *F. excelsior* trees of Moffat
621 Dale that are isolated from others by a distance of at least 250m (Figure 1), neither
622 produced a seed crop nor did they contribute to effective pollination of seeding
623 trees for the sampled reproductive season. On the basis of this observation, we
624 cannot reject the hypothesis that such isolated pasture trees are *living-dead*.
625 Similarly, only three of the 48 trees sampled locally have a high male reproductive
626 success and 26 of them contribute to effective pollination of fewer than two seeds

627 to none (Supplementary Table S1). However, whereas the evidence suggests that
628 most sampled trees have a low individual male reproductive success locally, we
629 cannot reject either the hypothesis that pollen from such trees effectively emigrated
630 to other forest remnants outside the sample area, as the presence of a large
631 component of immigrant pollination of either unknown origin or originated from
632 identified neighbouring sources would suggest (Figure 3).

633

634 Several ecological factors may have contributed to confer such an advantage to
635 long-distance pollination. Firstly, comparison of temporal variation of effective
636 pollen movement between mast seeding and non-masting years showed that in a
637 non-masting year, as is the case in the present study pollen-mediated gene flow
638 was favoured in a *F. excelsior* stand in southern England (FRAXIGEN 2005).
639 Furthermore, in such a situation, the small seed crops that were produced by a
640 number of trees (in particular, trees SCa34 and SCa38 displayed only one seed
641 branch with fewer than 10 seeds, Supplementary Table S1) may create a sampling
642 effect. Secondly, temporal variation in individual flowering phenology may greatly
643 affect mate availability. Indeed, Gerard *et al.* (2006) not only found that co-
644 flowering individuals were patchily distributed in space in a *F. angustifolia* and *F.*
645 *excelsior* hybrid zone, they also detected an asymmetry in male reproductive
646 success with early flowering trees participating more as pollen donors than late
647 flowering ones. A scenario where immigrant pollen would be preferentially available
648 during the period of stigma receptivity of seeding trees of Moffat Dale would also
649 favour long-distance pollination.

650

651 However, high levels of gene immigration are not necessarily sufficient to prevent
652 assortative mating and selfing (Gerard *et al.*, 2006). For gene flow to become an
653 efficient force counteracting the deleterious genetic effects of habitat

654 fragmentation, gene flow must not only be sustained at high levels across seasons,
655 but must also be qualitatively diverse. Here we find that efficient pollen immigration
656 allows for new and diverse genetic material to establish in the seed generation
657 (Bacles *et al.*, 2005). Such genetically diverse pollen pool composition may be
658 explained by the type of decay of pollen dispersal. Indeed, Klein *et al.* (2006)
659 demonstrated by means of simulation that fat-tailed dispersal kernels lead
660 asymptotically to a diverse propagule pool containing a balance of mixing of the
661 propagules of two sources and therefore that the diversity of the pollen pool of a
662 mother plant should increase with increased spatial isolation. Pollen dispersal
663 patterns observed for *F. excelsior* in Moffat Dale seem to corroborate such
664 theoretical findings. The majority of detected pollen dispersal was found between
665 near-neighbours at distances under 100m (Figure 2). However, not only were a
666 number of rare events detected among forest remnants at distances up to 2900m,
667 in proportions departing from random dispersal (Wilcoxon two-sided signed ranked
668 test, $p < 0.05$), but undetected events were also in the majority and may have
669 originated at much greater distances, suggesting a L-shaped pollen dispersal
670 kernel with a tail spreading over several kilometres and underlining the difficulty of
671 detecting long distance dispersal (Nathan, 2006).

672

673 Such pollen dispersal effective over long distances can be linked to landscape
674 features resulting from habitat disturbance. Indeed, in the Southern Uplands of
675 Scotland, deforestation and land use for pasture have greatly opened the barren
676 landscape that is regularly battered by strong winds. It is therefore likely that wind-
677 mediated pollen movement for *F. excelsior* has been facilitated by the modification
678 of the landscape in Moffat Dale. In particular, genetic connectivity of forest
679 remnants seems favoured by landscape openness and remnant size rather than by
680 geographic proximity. Indeed, although no seeding trees were sampled in two of

681 five remnants (SBa and Wa), fractional-likelihood paternity assignment shows that
682 their contribution to effective pollination of seeding trees of the other three extant
683 remnants (CDa, CMa and SCa) is higher than that of remnant CMa which is a
684 much smaller remnant of only 4 trees located in the bed of the river running
685 through an exposed and barren pasture in the Carrifran valley. Within remnant
686 pollination for CMa is indeed much lower than for other remnants (Figure 3),
687 highlighting the fact that remnants are smaller, but are spatially well connected to
688 neighbouring forest remnants tend to receive proportionally higher gene flow simply
689 because there are fewer potential local pollen donors (Sork and Smouse, 2006).

690

691 An important practical consequence of the high rate of pollination by immigrant
692 pollen is that the locally produced seed in Carrifran will contain genes sampled
693 from a wide geographic area around the valley. How the high rate of success of
694 immigrant pollen in the production of local seeds will ultimately affect the genetic
695 structure of *F. excelsior* remnants depends on how much of the pollen pool genetic
696 diversity is effectively carried into successive generations by established seedlings
697 that reach maturity. Natural regeneration in Moffat Dale has been severely limited
698 by continuous grazing pressure. Colonisation of mountain grasslands by *F.*
699 *excelsior* seedlings has been found to be connected to grazing activities, with
700 seedlings found preferentially in high layers of vegetation, in shaded and ungrazed
701 areas (Julien *et al.*, 2006). Pasture habitats in Moffat Dale may therefore be
702 unfavourable to seedling establishment. Thus, actual gene flow may be recruitment
703 limited rather than dispersal limited (Imbert and Lefèvre, 2003). In fact, comparison
704 of total gene flow estimated here from non-dispersed seeds with total gene flow
705 estimated from newly established seedlings in three *F. excelsior* remnants shows
706 that actual recruitment of genes carried by immigrant pollen is limited (Table 4).
707 Note that the ratio of potential to realised pollen-mediated gene flow is low, not

708 because there seems to be an advantage conferred to recruitment of local seeds
709 fertilised with local pollen but because the majority of establishing seedlings have
710 immigrated into the remnants (Bacles *et al.*, 2006). This indicates that, in cases
711 when seed dispersal is an important vector of long-distance dispersal, estimating
712 seed-mediated gene flow is essential to predicting landscape connectivity (Sork
713 and Smouse 2006).

714

715 In the Southern Uplands of Scotland and in other severely deforested landscapes,
716 conservation management aimed at sustainable forest restoration without human
717 intervention must move away from *conservation gardening* (Hobbs, 2007).

718 Predictive conservation necessitate better understanding of the evolution of
719 dispersal in a changing environment (Kokko and Lopez-Sepulcre, 2006) and an
720 appreciation of how population genetic processes operate in ecological space and
721 time. Bringing knowledge of contemporary gene flow among population remnants
722 generated from this study and others into conservation will ensure that the
723 evolutionary processes maintaining genetic connectivity and evolutionary potential
724 are restored at the landscape scale (Meagher, 2007).

725 **Acknowledgements**

726 This study was carried out between 2000 and 2003 with financial support from The
727 Scottish Forestry Trust (Registered Charity No. SC008465) to RAE in partial
728 fulfillment of CFEB's Doctorate of Philosophy of the University of Edinburgh which
729 was also supported by a Ph.D. studentship of the Natural Environment Research
730 Council. CFEB also gratefully acknowledges current funding from an Outgoing
731 International Fellowship of the European Union (Marie Curie Individual Actions,
732 contract CT-2005-9114 ADaPtE) for writing this article. We would like to thank, first
733 and foremost, AJ Lowe for use of sequencing facilities at CEH Edinburgh and
734 advice to CFEB during the completion of her Ph.D. Thanks are also due to S.
735 Cavers, J. Cottrell and R. Munro, for advice in the laboratory. Thank you to AC
736 Newton and The Wildwood Group of the Borders Forest Trust. We thank Tilhill
737 Forestry, J. Barker and C. Moffat for field access in Moffat Dale, especially during
738 the foot-&-mouth outbreak and M. Young for help with sample collection. S. Gerber
739 and E. Klein provided useful advice gene flow inference in FAMOZ. We are indebted
740 to J. Pemberton and R. Nathan and three anonymous reviewers for their insightful
741 comments that contributed to improving this article.

742 **Literature cited**

- 743 Aldrich PR, Hamrick JL (1998) Reproductive dominance of pasture trees in a
744 fragmented tropical forest mosaic. *Science* **28**: 103-105.
- 745 Amos W, Hoffman JI, Frodsham A, *et al.* (2007) Automated binning of
746 microsatellite alleles: problems and solutions. *Mol. Ecol. Notes* **7**: 10-14.
- 747 Araki H, Blouin MS (2005) Unbiased estimation of relative reproductive success of
748 different groups: evaluation and correction of bias caused by parentage
749 assignment errors. *Mol. Ecol.* **14**: 4097-4109.
- 750 Bacles CFE, Burczyk J, Lowe AJ, Ennos RA (2005) Historical and contemporary
751 mating patterns in remnant populations of the forest tree *Fraxinus excelsior* L.
752 *Evolution* **59**: 979-990.
- 753 Bacles CFE, Lowe AJ, Ennos RA (2006) Effective seed dispersal across a
754 fragmented landscape. *Science* **311**: 628.
- 755 Brachet S, Jubier MF, Richard M, Jung-Muller B, Frascaria-Lacoste N (1999) Rapid
756 identification of microsatellite loci using 5' anchored PCR in the common ash
757 *Fraxinus excelsior*. *Mol. Ecol.* **8**: 157-168.
- 758 Bittencourt JVM, Sebbenn AM (2007) Patterns of pollen and seed dispersal in a
759 small fragmented population of the wind-pollinated tree *Araucaria angustifolia* in
760 southern Brazil. *Heredity* **99**: 580-591.
- 761 Burczyk J, Adams WT, Moran GF, Griffin AR (2002) Complex patterns of mating
762 revealed in a *Eucalyptus regnans* seed orchard using allozyme markers and the
763 neighbourhood model. *Mol. Ecol.* **11**: 2379-2391.
- 764 Burczyk J, Chybicki IJ (2004) Cautions on direct gene flow estimation in plant
765 populations. *Evolution* **58**: 956-963.
- 766 Dakin EE, Avise JC (2004) Microsatellite null alleles in parentage analysis.
767 *Heredity* **93**: 504-509.

- 768 Dick CW (2001) Genetic rescue of remnant tropical trees by an alien pollinator.
769 *Proc. Roy. Soc. London B* **268**: 2391-2396.
- 770 Dow BD, Ashley MV (1998) High levels of gene flow revealed by paternity analysis
771 using microsatellites. **89**: 62-70
- 772 Ferrazzini D, Monteleone I, Belletti P (2007) Genetic variability and divergence
773 among italian populations of common ash (*Fraxinus excelsior* L.). *Ann. For. Sci.* **64**:
774 159-168.
- 775 FRAXIGEN (2005) *Ash species in Europe: biological characteristics and practical*
776 *guidelines for sustainable use* Oxford Forestry Institute, University of Oxford, U.K.
777 128 pp.
- 778 Gerard PR, Klein E, Austerlitz F, Fernandez-Manjarres JF, Frascaria-Lacoste N
779 (2006) Assortative mating and differential male mating success in an ash hybrid
780 zone population. *BMC Evol. Biol.* **6**:
- 781 Gerber S, Chabrier P, Kremer A (2003) FAMOZ: a software for parentage analysis
782 using dominant, codominant and uniparentally inherited markers. *Mol. Ecol. Notes*
783 **3**: 479-481.
- 784 Gerber S, Mariette S, Streiff R, Bodenes C, Kremer A (2000) Comparison of
785 microsatellites and amplified fragment length polymorphism markers for parentage
786 analysis. *Mol. Ecol.* **9**: 1037-1048.
- 787 Hadfield JD, Richardson DS, Burke T (2006) Towards unbiased parentage
788 assignment: combining genetic behavioural and spatial data in a Bayesian
789 framework. *Mol. Ecol.* **15**: 3715-3730.
- 790 Hamrick J (2004) Response of forest trees to global environmental changes. *For.*
791 *Ecol. Manage.* **197**: 323-335.
- 792 Hazlitt SL, Sigg DP, Eldridge MDB, Goldizen AW (2006) Restricted mating
793 dispersal and strong breeding group structure in a mid-sized marsupial mammal
794 (*Petrogale penicillata*). *Mol. Ecol.* **15**: 2997-3007.

- 795 Hebel I, Aldinger E, Haas R, *et al.* (2007) Pollen dispersal and pollen contamination
796 in an ash seed orchard (*Fraxinus excelsior* L.). *Allgemeine forst und jagdzeitung*
797 **178**: 44-49.
- 798 Heuertz M, Hausman F, Hardy OJ, Vendramin GG, Frascaria-Lacoste N,
799 Vekemans X (2004) Nuclear microsatellites reveal contrasting patterns of genetic
800 structure between western and southeastern European populations of the common
801 ash (*Fraxinus excelsior* L.). *Evolution* **58**: 976-988.
- 802 Heuertz M, Hausman F, Tsvetkov I, Frascaria-Lacoste N, Vekemans X (2001)
803 Assessment of genetic structure within and among Bulgarian populations of the
804 common ash (*Fraxinus excelsior* L.). *Mol. Ecol.* **10**: 1615-1623.
- 805 Hobbs R (2007) Managing plant populations in fragmented landscapes: restoration
806 or gardening? *Aus. J. Bot.* **55**: 371-374.
- 807 Hoffman JI, Amos W (2005) Microsatellite genotyping errors: detection
808 approaches, common sources and consequences for paternity exclusion. *Mol.*
809 *Ecol.* **14**: 599-612.
- 810 Imbert E, Lefèvre F (2003) Dispersal and gene flow of *Populus nigra* (Salicaceae)
811 along a dynamic river system. *J. Ecol.* **91**: 447-456.
- 812 Janzen DH (1986) Blurry catastrophes. *Oikos* **47**: 1-2.
- 813 Jones AG, Ardren WR (2003) Methods of parentage analyses in natural
814 populations. *Mol. Ecol.* **12**: 2511-2523.
- 815 Julien M-P, Alard D, Balent G (2006) Patterns of ash (*Fraxinus excelsior* L.)
816 colonization in mountain grasslands: the importance of management practices.
817 *Plant Ecology* **183**: 177-189.
- 818 Jump AS and Penuelas J (2006) Genetic effects of chronic habitat fragmentation in
819 a wind-pollinated tree. *Proc. Natl. Acad. Sci. USA* **103**: 8096-8100.
- 820 Klein E, Lavigne C, Gouyon P-H (2006) Mixing of propagules from discrete sources
821 at long distances: comparing a dispersal tail to an exponential. *BMC Ecol.* **6**: 3.

- 822 Kokko H, Lopez-Sepulcre A (2006) From individual dispersal to species range:
823 perspectives in a changing world. *Science* **313**: 789-790.
- 824 Lefort F, Brachet S, Frascaria-Lacoste N, Edwards KJ, Douglas GC (1999)
825 Identification and characterization of microsatellite loci in ash (*Fraxinus excelsior*
826 L.) and their conservation in the olive family (Oleaceae). *Mol. Ecol.* **8**: 1075-1092.
- 827 Lowe AJ, Boshier DH, Ward M, Bacles CFE, Navarro C (2005) Genetic resource
828 impacts of habitat loss and degradation: reconciling empirical evidence and
829 predicted theory for neotropical trees. *Heredity* **95**: 255-273.
- 830 Marshall TC, Slate J, Kruuk LEB, Pemberton J (1998) Statistical confidence for
831 likelihood-based paternity inference in natural populations. *Mol. Ecol.* **7**: 639-655.
- 832 Meagher TR (2007) Paternity analysis in a fragmented landscape. *Heredity* **99**:
833 563-564.
- 834 Morand ME, Brachet S, Rossignol P, Dufour J, Frascaria-Lacoste N (2002) A
835 generalized heterozygote deficiency assessed with microsatellites in French
836 common ash populations. *Mol. Ecol.* **11**: 377-385.
- 837 Morissey MB, Wilson AJ (2005) The potential costs of accounting for genotypic
838 errors in molecular parentage analyses. *Mol. Ecol.* **14**: 4111-4121.
- 839 Nathan R (2006) Long-distance dispersal of plants. *Science* **313**: 786-788.
- 840 Nielsen R, Mattila DK, Clapham PJ, Palsbøll PJ (2001) Statistical approaches to
841 paternity analysis in natural populations and applications to the north Atlantic
842 Humpback whale. *Genetics* **157**: 1673-1682.
- 843 O'Connell LM, Mosseler A, Rajora OP (2006) Impacts of forest fragmentation on
844 the mating system and genetic diversity of white spruce (*Picea glauca*) at the
845 landscape scale. *Heredity* **97**: 418-426.
- 846 Oddou-Muratorio S, Houot M-L, Demesure-Musch B, Austerlitz F (2003) Pollen
847 flow in the wild service tree, *Sorbus torminalis* (L.) Crantz. I. Evaluating the
848 paternity analysis procedure in continuous populations. *Mol. Ecol.* **12**: 3427-3439.

- 849 Ritland K (2002) Extensions of models for the estimation of mating systems using n
850 independent loci. *Heredity* **88**: 221-228.
- 851 Signorovitch J, Nielsen R (2002) PATRI-paternity inference using genetic data.
852 *Bioinformatics* **18**: 341-342.
- 853 Slate J, Marshall T, Pemberton J (2000) A retrospective assessment of the
854 accuracy of the paternity inference program CERVUS. *Mol. Ecol.* **9**: 801-808.
- 855 Slavov GT, Howe GT, Gyaourova AV, Birkes DS, Adams WT (2005) Estimating
856 pollen flow using SSR markers and paternity exclusion: accounting for mistyping.
857 *Mol. Ecol.* **14**: 3109-3121.
- 858 Smouse PE, Sork VL (2004) Measuring pollen flow in forest trees. *For. Ecol.*
859 *Manage.* **197**:21-38.
- 860 Sork VL, Davis FW, Smouse PE, Apsit PE, Dyer RJ, Fernandez, JF *et al.* (2002)
861 Pollen movement in declining populations of California Valley oak, *Quercus lobata*:
862 where have all the fathers gone? *Mol. Ecol.* **11**: 1657-1668.
- 863 Sork VL, Nason JD, Campbell DR, Fernandez JF (1999) Landscape approaches to
864 historical and contemporary gene flow in plants. *Trends Ecol. Evol.* **14**: 219-224.
- 865 Sork VL, Smouse PE (2006) Genetic analysis of landscape connectivity in tree
866 populations. *Landsc. Ecol.* **21**: 821-836.
- 867 Wagner AP, Creel S, Kalinowski ST (2006) Estimating relatedness and
868 relationships using microsatellite loci with null alleles. *Heredity* **97**: 336-345
- 869 White GM, Boshier DH, Powell W (2002) Increased pollen flow counteracts
870 fragmentation in a tropical dry forest: An example from *Swietenia humilis* Zuccarini.
871 *Proc Natl. Acad. Sci. USA* **99**: 2038-2042.
- 872 Whitlock MC, McCauley DE (1999) Indirect measures of gene flow and migration:
873 F_{st} different $1/(4Nm+1)$. *Heredity* **82**: 117-125.
- 874 Young A, Boyle T, Brown AHD (1996) The population genetic consequences of
875 habitat fragmentation for plants. *Trends. Ecol. Evol.* **11**: 413-418.

876 **Titles and legends of figures**

877

878 **Figure 1** Distribution of *Fraxinus excelsior* mature trees in Moffat Dale remnants.

879 Each dot represents a tree. A) Mature trees grow in small forest remnants,

880 confined to steep slopes (elevations are given in meters) along streams

881 (highlighted in dark lines). An exhaustive sampling and mapping was performed in

882 remnants CMa ($N=4$), CDa ($N=30$) and SCa ($N=12$) while in two larger remnants

883 SBa and Wa which include approximately 50 mature trees each, 20 individuals

884 were sampled throughout each of them as potential sources of immigrant pollen

885 flowing into remnants CMa, CDa and SCa. Two lone trees of the Carrifran Burn

886 (Labelled A and B) were also sampled. B to F) Close up of spatial distribution of

887 individuals sampled in remnant CDa, CMa, SCa, SBa and Wa respectively. In CDa,

888 CMa, and SCa, all individuals producing fruits in 2000 are represented by a star: 30

889 seeds were collected throughout the tree canopy from 11, 2 and 6 trees in CDa,

890 CMa and SCa respectively. The background map is a section of Ordnance Survey

891 product Land-line.Plus-nt11 © Crown copyright Ordnance Survey. An EDINA

892 digimap / JISC supplied service. © Figure 1 was originally published in Bacles *et al.*

893 (2005) reprinted with kind permission from *Evolution* (Blackwell Publishing)..

894

895

896 **Figure 2** Comparison of the frequency distribution of possible (white) and detected

897 (black) effective pollen dispersal events for *Fraxinus excelsior* in relation to the

898 location of pollen donors within the fragmented landscape of Moffat Dale. 2a)

899 Detection within three censused remnants. Possible dispersal distances were

900 estimated from Euclidian distances between 141 assignable seeds, spatially

901 located on 19 mother trees, and all 48 *F. excelsior* candidate pollen parents

902 sampled in remnants CDa, CMa and SCa of Moffat Dale (Figure 1). Detected

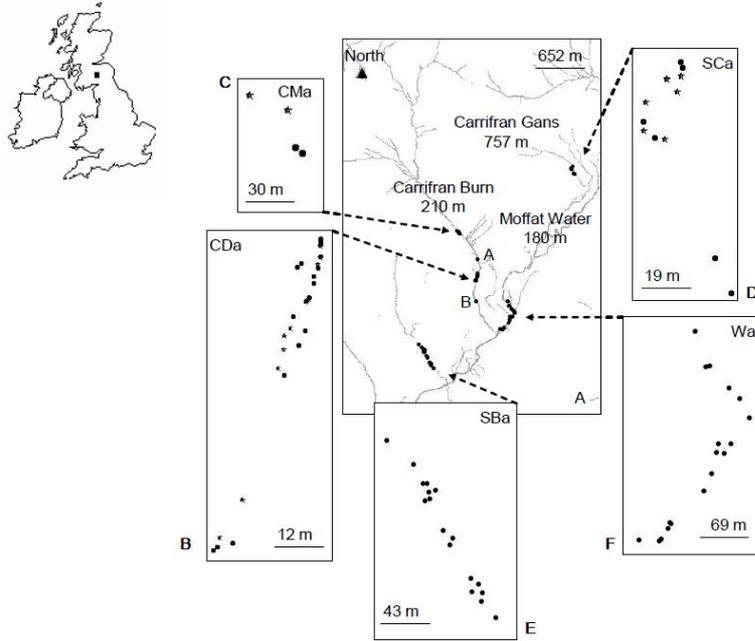
903 pollen dispersal distances were estimated from Euclidian distances between the
904 141 assignable seeds and their likely father when identified by means of maximum-
905 likelihood paternity analysis in FAMOZ (Gerber et al. 2003, Table 2). 2b) Detection
906 including neighbouring sources of immigrant pollen. Possible dispersal distances
907 were estimated from Euclidian distances between 163 assignable seeds, spatially
908 located on 19 mother trees, and all 88 *F. excelsior* candidate pollen parents
909 sampled in all five remnants of Moffat Dale (Figure 1). Detected pollen dispersal
910 distances were estimated from Euclidian distances between the 163 assignable
911 seeds and their likely father when identified by means of maximum-likelihood
912 paternity analysis in FAMOZ (Gerber et al. 2003, Table 3). In both situations, when
913 more than one likely father was identified (unresolved assignment), a fraction of the
914 seed was assigned to all likely fathers evenly and proportionally to the number of
915 likely fathers (Supplementary Table S1 and S2). Distance distributions of detected
916 pollen dispersal were found to differ significantly from random dispersal (Wilcoxon
917 two-sided signed rank test, p -value <0.05 for both $n=141$ and $n=163$).

918

919 **Figure 3** Schematic map of pollen-mediated genetic exchange among three
920 *Fraxinus excelsior* forest remnants varying in their population size, density and
921 degree of spatial isolation to other forest remnants in the mosaic landscape of
922 Moffat Dale. Estimates of gene movement within remnants (continuous white
923 arrows), of gene flow among remnants (continuous black arrows) and of gene
924 immigration from external sources (dashed white arrows) are based on results of
925 ML paternity analysis performed in FAMOZ (Gerber et al. 2003) 422 seeds sampled
926 from all 19 seeding trees in remnants CDa ($N_{trees}=30$), CMa ($N_{trees}=4$) and SCa
927 ($N_{trees}=12$), considering all 48 trees occurring within the landscape, including two
928 isolated trees (A and B, Figure 1) as potential pollen donors (Table 2). Relevant
929 potential geographic barriers to gene flow among remnants are highlighted:

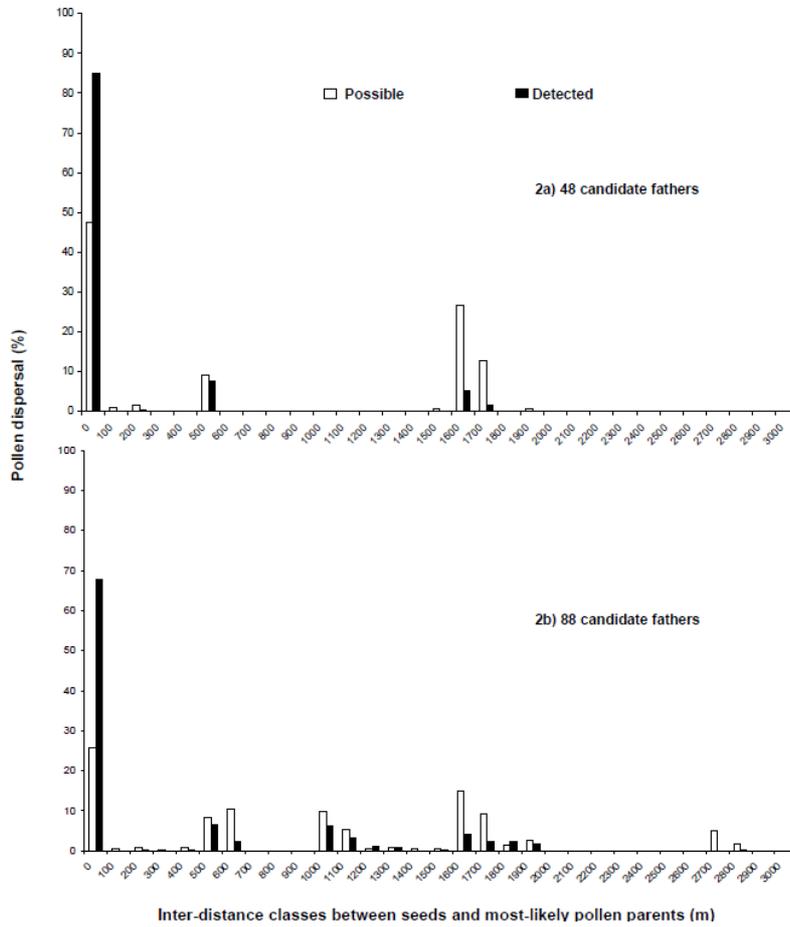
930 Remnants CDa and CMa are located in close proximity (600 m) at the bottom of
931 the bare and open valley while remnant SCa is most isolated over a ridge (dashed
932 white rectangle) located about 1700 m away and surrounded by a dense closed
933 conifer plantation (continuous black square).

934 FIGURE 1



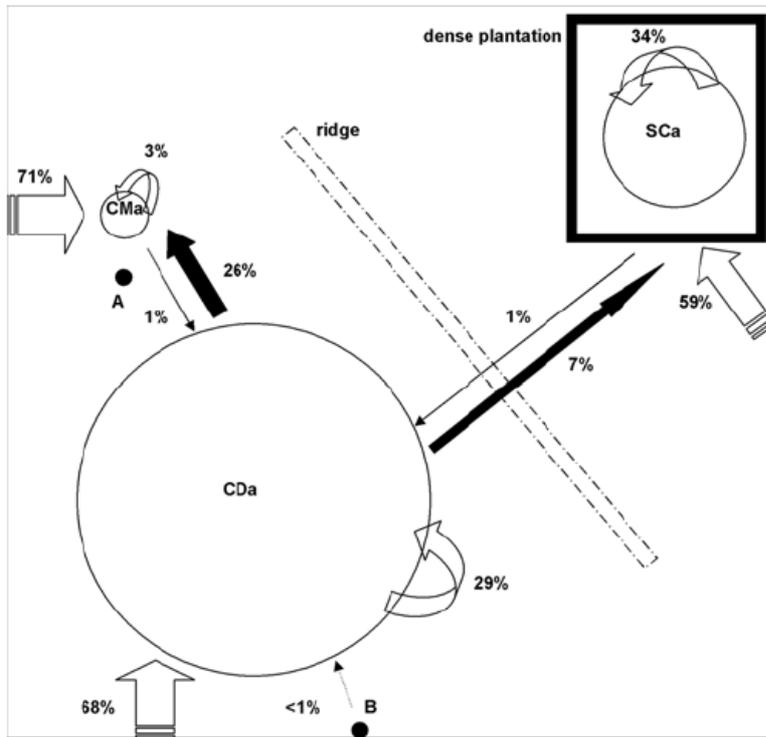
935

936 FIGURE 2



937

938 FIGURE3



939

940 **Table 1.** Estimates of genotyping error and paternity exclusion probabilities (PEP) at each of five *Fraxinus excelsior* microsatellite loci and overall,
 941 computed in CERVUS 2.0 (Marshall et al. 1998) and based on genotyping of 483 seeds sampled from 19 mothers and of 88 trees with no transformation
 942 of genotypes (RAW), with binning of rare alleles (BIN), with binning of rare alleles and inclusion of a generalised null allele (NULL). Detailed procedures for
 943 allele binning and null allele inclusion are given in material and methods section. Genotyping error estimates based on comparison of the genotype of
 944 the 483 seeds with that of their mother. The number of mismatching seed genotypes is given (N_{mismatch}) but the number of comparisons may be under
 945 483 when values are missing ($N_{\text{comparison}}$). Lowest estimates of genotyping error for each locus are highlighted in bold.

Dataset	RAW			BIN			BINNULL		
	$N_{\text{mismatch}}/N_{\text{comparison}}$	error	PEP	$N_{\text{mismatch}}/N_{\text{comparison}}$	error	PEP	$N_{\text{mismatch}}/N_{\text{comparison}}$	error	PEP
3.1	129/431	0.2847	0.692	118/431	0.2679	0.680	69/483	0.1499	0.653
FEMSATL5	132/434	0.2237	0.808	102/434	0.2133	0.713	39/483	0.0762	0.685
1.19	13/469	0.0332	0.594	9/469	0.0292	0.504	12/483	0.0289	0.606
M2.30B	26/460	0.0378	0.855	23/460	0.0347	0.837	45/483	0.0641	0.841
FEMSATL2	34/481	0.0463	0.864	32/481	0.0443	0.857	34/483	0.0457	0.870
Overall (excl. 3.1 and 5)	-	0.0391	0.992	-	0.0361	0.988	-	0.0462	0.992
Overall (all)	-	0.1251	0.999	-	0.1179	0.999	-	0.0732	0.999

946 **Table 2.** Comparison of global results of paternity analysis of 422 *Fraxinus excelsior* seeds sampled from 19 mother trees and 48 candidate fathers and
 947 of their translation into percentage of apparent pollen-mediated gene flow into three forest remnants of the Moffat Dale catchment for a range of
 948 paternity assignment methods and of microsatellite genotype transformations. Simple exclusion (SE) and maximum-likelihood (ML) paternity analyses
 949 were conducted in FAMOZ (Gerber *et al.*, 2003). Lod of the odds score (LOD) thresholds (TF) for ML paternity assignment and Type I error of false
 950 rejection and Type II error of false assignment were determined by means of 50 000 simulations in FAMOZ. Both SE and ML methods were applied to
 951 seeds genotyped at microsatellites FEMSATL2 (Brachet *et al.*, 1999), M2.30B and 1.19 (Lefort *et al.*, 1999) that displayed a multilocus genotype
 952 compatible with the multilocus genotype of their mother, with no further transformation (RAW), after allele binning at all three loci (BIN), after additional
 953 inclusion of a generalised null allele at locus 1.19 (BIN3NULL1). Results highlighted in bold were retained for detailed analysis of genetic connectivity
 954 among remnants.

Dataset	Paternity assignment method	TF	Type I error	Type II error	N_{seed} Excluded LOD=0	N_{seed} Not assigned	N_{seed} Assigned, resolved	N_{seed} Assigned, unresolved	Apparent pollen flow LOD≤TF
						0<LOD<TF	LOD>TF, unique	LOD>TF, multiple	
RAW		0	-	-	219	-	120	83	52%
BIN	SE	0	-	-	183	-	125	114	43%
BIN3NULL1		0	-	-	224	-	119	79	53%
RAW		3.20	<0.10	<0.13	219	68	102	33	68%
BIN	ML	2.50	<0.07	<0.28	183	41	135	63	53%
BIN3NULL1		2.90	<0.05	<0.22	224	57	92	49	67%

955 **Table 3.** Comparison of fractional and maximum likelihood estimation of the contribution of the five *Fraxinus excelsior* forest remnants of Moffat Dale to
 956 effective pollination of seeds sampled in three of them. Results are based on the 422 seeds sampled in remnants CDa, CMa and SCa which displayed
 957 a genotype compatible with that of their mother, transformed to minimise genotyping error. †20 trees were genotyped in each of SBa and Wa as an
 958 indication of potential local sources of immigrant pollen. The most-likely number of seeds sired in each remnant was determined by means of maximum-
 959 likelihood paternity analysis in FAMOZ (Gerber *et al.*, 2003) considering all 88 trees genotyped and assigning paternity at a threshold T_F of 2.92 (Type I
 960 error <0.05, Type II error <0.28). The fractional-likelihood paternity analysis was computed in PATRI (Signorovitch and Nielsen, 2002). The posterior
 961 expectation of the number of seeds fertilised by trees from each of five forest remnants was estimated for a given population size (N) of 88 (the number
 962 of trees genotyped), of 150 (approximating the total number of trees occurring in Moffat Dale) and modelled as a uniform function of 100-500
 963 individuals.

		Fractional likelihood assignment in PATRI			Maximum likelihood assignment in FAMOZ	
		Posterior expectations of the number of sampled seeds			Most likely number of seed sired	
		Prior N	88	150	Uniform [100-500]	88
Remnant	CDa	88	56	23	91	
	CMa	3	2	1	4	
	SCa	32	28	19	35	
Gene flow[†]	SBa	15	11	6	18	
	Wa	8	7	4	15	
All		146	104	53	163	

964 **Table 4.** Comparison of pollen and total gene flow estimates from maximum-likelihood (ML) paternity analysis of non-dispersed seeds with estimates
 965 from ML parent-pair analysis of established seedlings in three *Fraxinus excelsior* forest remnants of Moffat Dale. Forest remnants CDa, CMa and SCa
 966 were exhaustively sampled for 48 adult trees which were all considered as potential pollen donors and seeds were sampled from all trees producing
 967 fruits, respectively, 11, 2 and 6 trees in CDa, CMa and SCa. [#]sample size given as number of seed families/total number of seeds. ^{*}Results of ML
 968 parent-pair analysis of seedlings establishing in *F. excelsior* remnants CDa, CMa and SCa were previously reported in Bacles *et al.* (2006). [§]See
 969 Equation 1 estimating total gene flow from progeny arrays (T_p) and where A is the number of local seeds fertilised by immigrant pollen and C is the total
 970 number of seeds sampled [†]See Equation 2 estimating total gene flow from established seedlings (T'_s) and where A' is the number of seeds fertilised by
 971 immigrant pollen, B' is the number of immigrating seeds and C' is the total number of established seedlings sampled.

	progeny arrays: ML paternity analysis <i>ante</i> seed dispersal					established seedlings: ML parent-pair analysis <i>post</i> establishment						
	origin of pollen (N_{seed})			gene flow estimates		origin of pollen and seed (N_{seed})				gene flow estimates		
	sample size [#]	local seed pollen	foreign pollen	pollen flow	total gene flow via pollen	sample size	local seed pollen	foreign pollen	foreign seed	potential pollen flow	realised pollen flow	total gene flow
remnant	C^{\S}		A^{\S}		T_p^{\S}	C'^{\dagger}		A'^{\dagger}	B'^{\dagger}	<i>local seed</i>		T'_s^{\dagger}
CDa	11/282	81	201	71%	35%	20	3	7	10	70%	17.5%	67.5%
CMa	2/32	1	31	97%	48%	20	0	5	15	100%	12.5%	87.5%
SCa	6/108	37	71	66%	33%	20	2	7	11	78%	17.5%	72.5%

972

973 **Supplementary Table S1.** Individual contribution of *Fraxinus excelsior* trees to effective pollination of non-dispersed seeds sampled in three forest
 974 remnants of the Moffat Dale catchment. Results are based on the 422 seeds sampled in remnants CDa, CMa and SCa which displayed a genotype
 975 compatible with that of their mother, transformed to minimise genotyping error, considering all 48 sires growing in the three remnants. The most-likely
 976 number of seeds sired by individual trees was determined by means of maximum-likelihood paternity analysis in FAMOZ (Gerber et al. 2003) considering
 977 48 candidates genotyped and assigning paternity at a threshold T_F of 2.90 (Type I error <0.05, Type II error <0.22). When more than one likely father
 978 was identified (unresolved assignment), a fraction of the seed was assigned to all likely fathers evenly and proportionally to the number of likely fathers.
 979

		MOTHER TREE																						
		CDa										CMa			SCa									
SIRE		102	105	108	111	112	118	123	125	126	129	130	CDa all	24	26	CMa all	33	34	35	36	38	41	SCa all	All
	<i>n</i>	22	29	23	29	24	21	27	26	23	30	29	283	20	11	31	22	2	28	24	3	29	108	422
CDa	19					0.33	1.00		9.33	8.00	1.00		19.67											19.67
	130	1.00	1.00							4.00		1.00	7.00											7.00
	107			1.00	2.00							3.00	6.00											6.00
	109		0.33	1.33		2.00	1.33	1.00					6.00											6.00
	114		1.00			1.00		1.00				2.33	5.33							0.50			0.50	5.83
	110		3.00		0.75	1.00				1.00			5.75											5.75
	113								1.00		0.50		1.50	0.50		0.50	1.00		1.00	0.67			2.67	4.67
	116		1.00			1.00						1.00	3.00	1.50		1.50								4.50
	119	0.33	1.00						1.00			2.00	4.33											4.33
	115	0.33				0.33	1.00		0.83		0.50		3.00	1.00		1.00								4.00
	125		0.33	0.33	1.00		1.33	1.00					4.00											4.00
	120													1.00	2.00	3.00								3.00
	101				0.25								0.25		0.50	0.50			1.00	1.00			2.00	2.75
	111				1.50							1.00	2.50											2.50
	129		0.39			1.00	1.00						2.39											2.39
	104		0.25								0.83		1.08	0.50		0.50				0.50			0.50	2.08
	118		0.14										0.14	0.50		0.50				0.17		1.00	1.17	1.81

36																			0.50	0.50	0.50									
31																			0.50	0.50	0.50									
34		0.25																	0.25	0.25	0.25									
35																														
SCa all	0.39		0.50																2.00	2.89			10.00	2.00	11.00	11.67	0.00	2.00	36.67	39.56
B	0.25																		0.33	0.58										0.58
A																														
All sires	2	13	3	8	9	7	8	13	9	12	4	88	6	3	9	11	2	13	15	0	3	44	141							

980

981 **Supplementary Table S2.** Individual contribution of *Fraxinus excelsior* trees to effective pollination of non-dispersed seeds sampled in three forest
 982 remnants of the Moffat Dale catchment. Results are based on the 422 seeds sampled in remnants CDa, CMa and SCa which displayed a genotype
 983 compatible with that of their mother, transformed to minimise genotyping error, considering 88 sires. 20 trees were genotyped in each of remnant SBa
 984 and Wa as an indication of potential local sources of immigrant pollen. The most-likely number of seeds sired by individual trees was determined by
 985 means of maximum-likelihood paternity analysis in FAMOZ (Gerber et al. 2003) considering all 88 trees genotyped and assigning paternity at a threshold
 986 TF of 2.92 (Type I error <0.05, Type II error <0..28). When more than one likely father was identified (unresolved assignment), a fraction of the seed was
 987 assigned to all likely fathers evenly and proportionally to the number of likely fathers.
 988

SIRE	MOTHER TREE																						All
	CDa											CMa				SCa							
	102	105	108	111	112	118	123	125	126	129	130	CDa all	24	26	CMa all	33	34	35	36	38	41	SCa all	
<i>n</i>	22	29	23	29	24	21	27	26	23	30	29	283	20	11	31	22	2	28	24	3	29	108	422
19					0.25	1.00		9.33	8.00	1.00		19.58											19.58
130	1.00	1.00						4.00		1.00		7.00											7.00
109		0.33	1.33		2.00	1.33	1.00					6.00											6.00
110		3.00		0.70	1.00				1.00			5.70											5.70
107			1.00	1.50						2.50		5.00											5.00
114		0.50			0.50		1.00			2.25		4.25							0.50			0.50	4.75
113								1.00	0.50			1.50	0.50		0.50	1.00		1.00	0.50			2.50	4.50
119	0.33	1.00						1.00		2.00		4.33											4.33
125		0.33	0.33	1.00		1.33	1.00					4.00											4.00
115	0.33				0.25	1.00		0.83	0.50			2.92	1.00		1.00								3.92
120													1.00	2.00	3.00								3.00
116		0.67			0.50					1.00		2.17	0.75		0.75								2.92
101				0.20								0.20		0.50	0.50			1.00	1.00			2.00	2.70
111				1.50						1.00		2.50											2.50
118		0.14										0.14	0.50		0.50				0.17		1.00	1.17	1.81
129		0.24			1.00	0.50						1.74											1.74

	104	0.10									0.75	0.85	0.50	0.50				0.33	0.33	1.68			
	128			0.50	1.00							1.50								1.50			
	102	0.15		1.20								1.35								1.35			
	112	0.33	1.00									1.33								1.33			
	127							0.33	0.50	0.50		1.33								1.33			
	105	0.48	0.33					0.33				1.14						0.17	0.17	1.31			
	126	1.00		0.20								1.20								1.20			
	121	0.14										0.14	0.50	0.50				0.17	0.17	0.81			
	124	0.67										0.67								0.67			
	122	0.14										0.14						0.17	0.17	0.31			
	103																						
	108																						
	117																						
	123																						
	CDa all	2.00	10.90	3.00	6.80	6.50	5.83	8.00	11.67	9.00	11.00	2.00	76.70	4.75	2.50	7.25	1.00	2.00	3.00	1.00	7.00	90.95	
CMa	26			0.50	1.25							1.33	3.08								3.08		
	24												0.25	0.50	0.75						0.75		
	27																						
	28																						
	CMa all			0.50	1.25							1.33	3.08	0.25	0.50	0.75					3.83		
SCa	32																3.50	1.00	4.00	5.00	13.50	13.50	
	37																0.50		7.00	4.00	1.00	12.50	12.50
	42																1.75			1.00		2.75	2.75
	33																1.75			0.50		2.25	2.25
	38		0.14								0.50	0.64						0.50		0.17	0.50	1.17	1.81
	41																			0.50		0.50	0.50
	36																			0.50		0.50	0.50

	8																			
	13																			
	14																			
	19																			
	SBa all	1.67	3.47	3.00	2.20	1.25	1.50	1.00	1.00	0.75	15.83	1.00	1.00	2.00	0.50				0.50	18.33
Wa	11													0.50	1.00	0.50	2.00		2.00	
	15		1.00								1.00				1.00		1.00		2.00	
	16	0.33			0.50		0.50							0.50				0.50	1.83	
	13								0.33	0.33						1.00	1.00		1.33	
	12												1.25				1.25		1.25	
	19												1.25				1.25		1.25	
	4														1.00		1.00		1.00	
	6								1.00	1.00									1.00	
	14											1.00	1.00						1.00	
	18										1.00		1.00						1.00	
	1								0.50	0.50									0.50	
	7															0.33		0.33	0.33	
	3		0.10									0.10							0.10	
	8		0.10									0.10							0.10	
	17		0.10									0.10							0.10	
	2																			
5																				
9																				
10																				
20																				

Published in Heredity 101 (2008) : 368-380.

Wa all	0.33	1.30		0.50		0.50					1.83	4.47	1.00	1.00	2.00	3.00	0.50		3.33	1.50	8.33	14.80	
SBa+Wa	2	5	3	3	1	2	1	0	1	1	2	20	2	2	4	4	1	0	3	0	2	9	33
All sires	4	16	6	10	9	8	9	13	10	12	5	102	7	5	12	12	2	13	18	0	4	49	163

989
990